

Applying a water-energy-food nexus approach to seafood products from the European Atlantic area

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Abstract

By 2050, food production's environmental impacts are projected to double without intervention. Crucial changes in dietary habits are needed, and seafood can be pivotal. The focus in calculating burdens using life cycle assessment (LCA) indicators has shifted towards exploring interconnections within the Water-Energy-Food (WEF) nexus. Addressing this evolution, this manuscript has applied an innovative methodology to calculate a WEF nexus index (WEFni). This index seamlessly integrates both environmental and nutritional profiles across diverse case studies involving fisheries, aquaculture, and processing production systems in the European Atlantic area. The results showed that when it comes to fishing, purse seine fishing obtained the highest score (99%). For the aquaculture activities, seafood (>73%) obtained better results than fish farming. About processing treatments, freezing (>79%) has lower environmental loads than canning. In addition, a sensitivity analysis was carried out according to a caloric- and edible-based functional unit as well as the modification of the weighting factors assigned to each footprint, in which important variations were reported in the WEFni of most case studies. Consequently,

41 although the methodological guidelines supporting the WEFni have great potential to serve as an
42 ecolabeling tool, certain aspects need to be re-evaluated. On the other hand, it could be interesting
43 to include the potential biodiversity loss, as well as other socioeconomic indicators, with the
44 purpose of considering a more sustainability criteria. Finally, the contribution of this paper to the
45 food industry literature is of paramount importance, as it represents a first step towards holistic
46 assessments and easy-to-understand ecolabelling processes that can promote sustainable
47 production and consumption.

48 **Keywords**

49 Fisheries; Aquaculture; Life cycle assessment; Environmental footprints; Nutritional index;
50 Ecolabel.

51 **Highlights**

- 52 • A joint analysis of environmental burdens and nutritional profiles was carried out.
- 53 • The functional unit and weighting factors were modified in a sensitivity analysis.
- 54 • WEFni revealed: purse seine high score, shellfish outperformed fish farming, and
55 freezing had lower burdens.
- 56 • The procedure followed has potential applicability to other products.
- 57 • Further indicators are needed if a holistic analysis of sustainability is to be made.

58 **Acronyms**

59	CC	Climate change
60	CF	Carbon footprint
61	CI	Caloric intake
62	EF	Energy footprint
63	EoL	End of life
64	EP	Edible portion
65	FU	Functional unit
66	FU_{ce}	Caloric- and edible-based functional unit
67	HORECA	Hotels, restaurants and catering
68	LCA	Life cycle assessment
69	LCI	Life cycle inventory
70	LCIA	Life cycle impact assessment
71	NF	Nutritional Footprint
72	PEF	Product environmental footprint
73	PEFCR	Product environmental footprint category rules
74	RSI	Reference seafood intake
75	SM	Supplementary material
76	WEF	Water-Energy-Food
77	WEFni	Water-Energy-Food nexus index

79 **1. Introduction**

80 The food sector integrates a variety of activities such as agriculture, livestock, fisheries,
81 aquaculture, processing and distribution. Over time, food production has shifted towards intensive
82 systems characterised by higher yields to meet an intense food demand (Cucurachi et al., 2019).
83 This has also meant that the food system is related to multiple environmental impacts, such as
84 nutrient pollution (generating acidification or eutrophication from the use of fertilisers),
85 greenhouse gases (GHGs) emissions, land use, as well as water scarcity (Clark et al., 2019). At
86 the same time, food production activities are being affected by climate change (CC). This fact is
87 particularly significant in the case of agriculture, aquaculture and fisheries due to the loss of fertile
88 lands, redistribution of habitats or increasingly frequent extreme weather events (Barange et al.,
89 2018). It has been observed that CC impacts are unevenly distributed depending on the economic
90 activity or social stratum affected. For instance, model projections concluded that fisheries would
91 be more harmed in terms of production level compared to agricultural activities, apart from the
92 fact that the population of lower socio-economic status would experience these changes with
93 greater severity (Cinner et al., 2022).

94 A combination of measures need to be implemented considering that both the demographic and
95 income levels in most countries will raise and the food system will almost double its
96 environmental impacts by 2050, if the combined action of a number of measures is not carried
97 out, including a shift towards a less meat-based diet, the application of technological
98 breakthroughs, and a reduction of food loss and waste (Springmann et al., 2018). Therefore, food
99 system transformation is key to achieve food security, strengthen the management of natural
100 resources, while providing food in a sustainable manner to cope with CC consequences (Bogard
101 et al., 2019). In this regard, blue foods (i.e., animal, plants and algae harvested or farmed from
102 freshwater and marine environments) (Crona et al., 2023), are presented as potential candidates
103 for nourishing humankind at low environmental burdens on ecosystems (Gephart et al., 2021). In
104 addition, seafood products stand out due to their excellent nutritional profiles, being rich in
105 essential fatty acids (e.g. n-3), proteins (e.g. essential aminoacids), some micronutrients (e.g. Fe,
106 Ca, Zn, Se, I), and vitamins (e.g., A, C, D, E) (Golden et al., 2021). Thus, the consumption of
107 seafood products contributes to a healthy diet and offers a unique perspective to address the global
108 health “triple burden” issue of malnutrition: obesity, undernutrition, and micronutrient
109 deficiencies (Obiero et al., 2019).

110 Seafood products have also an important socio-economic relevance, being greatly present in the
111 diet of the European Atlantic region (i.e. Spain, Portugal, France, Ireland and United Kingdom),
112 with an average intake of 38.9 kg per capita, a significant higher consumption compared to the
113 European Union (EU) average (23.3 kg per capita) (EUMOFA, 2023). In terms of production, the
114 Atlantic area is recognised as one of the main exporters at the European level, reaching
115 manufacturing levels of 2.6 million tonnes of seafood in 2018 (Laso et al., 2022). These figures
116 are expected to raise over time, since it is estimated that global seafood production (from fisheries
117 and aquaculture) will reach 202 million tons by 2030 and aquaculture production will overtake
118 capture fisheries by 2027 (FAO, 2022).

119 In view of the environmental challenges that the seafood sector faces and its importance in a
120 socio-economic context, it needs a cross-sectional change to become a more resilient food source.
121 To achieve this goal, the life cycle assessment (LCA) methodology emerges as an accepted
122 analytical framework for assessing environmental impacts related to food production, and

123 proposing mitigation strategies (Ruiz-Salmón et al., 2021a). The application of the LCA
124 methodology to seafood products has generated a great deal of interest in recent years, with about
125 90% of the LCA studies published in the last decade being reported (Ruiz-Salmón et al., 2021b).
126 Based on the LCA concept, a specific guide for estimating the environmental footprint of marine
127 fish products has been developed recently (Marine Fish PEFCR, 2022). This is related to the
128 initiative of the Product Environmental Footprint (PEF) led by the European Commission
129 (European Commission, 2013), which recommends the use of a set of indicators to create a single
130 market of green products (European Commission, 2018).

131 However, the use of environmental indicators that focus on a single problem in an individual way
132 hampers the goal of achieving an integrated resource management (Purwanto et al., 2021). This
133 is because such indicators often overlook synergies and trade-offs between water demand, energy
134 requirements, and food provision (Cansino-Loeza et al., 2022). Likewise, the feedback loop
135 between these drivers of water, energy and food scarcity, which constitutes a triple nexus,
136 promotes a situation of competition for an increasingly limited number of basic resources in many
137 regions of the world (Fernández-Ríos et al., 2021). To highlight this issue, the Water-Energy-
138 Food (WEF) nexus concept was developed, providing a tool to balance these elements through a
139 holistic approach (Simpson and Jewitt, 2019). More specifically, for the seafood sector, the WEF
140 nexus index (WEFni) was proposed by Entrena-Barbero et al. (2023a) following an integrative
141 perspective, aggregating a set of indicators in a single value and reflecting different issues related
142 to the seafood sector as a whole.

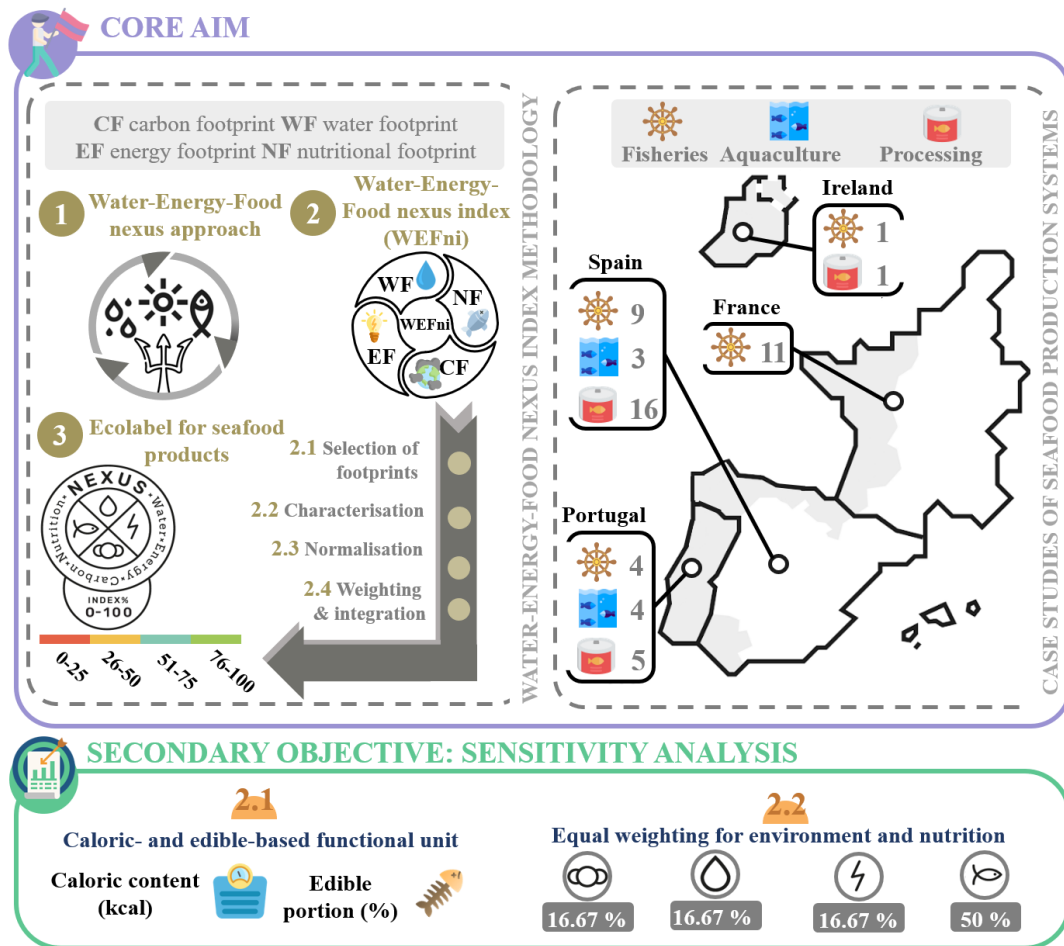
143 As a driver for raising awareness of such composite indexes, ecolabels are useful means of
144 communication between producers and consumers (Giacomarra et al., 2021) that are increasingly
145 concerned about sustainable consumption and production (Sigurdsson et al., 2022). Despite
146 decades of academic research and practice using ecolabels as a way of sustainability signalling,
147 over 60% of consumers still find sustainable food choices difficult to identify (International Food
148 Information Council, 2020). This is due to, among other reasons, the large number of ecolabels
149 available in the market (more than 400 worldwide, being 50 applicable to seafood products) and
150 the lack of a robust regulatory framework (Ecolabel Index, 2023). Therefore, a flood of over-
151 information is occurring in an attempt to advise consumers. However, only scientifically based
152 ecolabels and recommendations can generate consumer confidence and security by enabling
153 ‘green’ and conscious choices when buying seafood products. In this view, the application of the
154 WEFni in the form of an ecolabel will constitute a reliable ecolabelling scheme to promote an
155 environmental sustainability throughout the seafood value chain and to raise awareness on this
156 index.

157 In this context, the novelty and main goal of this study was to use for the first time the
158 methodological guidelines for the estimation of the WEFni elaborated by Entrena-Barbero et al.
159 (2023a) to a series of case studies in the European Atlantic area to analyse trends and hotspots of
160 both the single indicators (carbon footprint (CF), water footprint (WF), energy footprint (EF),
161 nutritional footprint (NF)), as well as the composite index proposed (i.e. WEFni). After this,
162 strengths and weaknesses of the WEFni methodology were identified with the aim of proposing
163 improvements in further iterations. Along with the main goal, a secondary objective was defined:
164 to carry out a sensitivity analysis of the results obtained by varying certain methodological aspects
165 of the method, in particular the functional unit (FU) and weighting factors. These guidelines were
166 configured within the framework of the NEPTUNUS project (Laso et al., 2022) and their
167 application allows benchmarking seafood products. The study will contribute to the development
168 of a useful framework for the estimation of the environmental, energy and nutritional profiles of

169 seafood products from fisheries or aquaculture production systems, or which have undergone a
 170 type of processing.

171 **2. Methodology**

172 The main features of the methodological procedure followed in this paper are summarised in
 173 **Figure 1**. The methodological guidelines to calculate the WEFni are described in **Section 2.1**,
 174 while the sensitivity analysis is presented in **Section 2.2**.



175

176 **Figure 1.** Main features of the methodological procedure.

177 **2.1. Water-Energy-Food nexus index methodological guidelines**

178 The WEFni methodological guidelines, proposed by Entrena-Barbero et al. (2023a), and
 179 underpinned by the international standards ISO 14040 (ISO, 2006a) and ISO 14044 (ISO, 2006b),
 180 proposes a common roadmap for the elaboration of LCA studies applied to seafood products,
 181 establishing aspects such as FU, system boundaries, allocation rules for multi-functional systems,
 182 minimum life cycle inventory (LCI) data required or end of life (EoL) modelling, among others.
 183 They also select the life cycle impact assessment (LCIA) methods used to estimate the following
 184 environmental footprints: CF, WF and EF. Moreover, they also propose a nutritional footprint in
 185 the form of a modified Nutrient Rich Food (NRF) index for the specific case of seafood products,
 186 having considered 12 nutrients to promote and 2 nutrients to limit (NRF12.2). Finally, they
 187 propose a process of normalisation (from 0 to 1) and weighting (25% for each footprint) to
 188 integrate the four indicators into a dimensionless single value (i.e., the WEFni), which varies from

189 0 to 100. The final value was represented in the form of front-of-pack ecolabel to enable a
 190 communication logo that can be applied by producers and perceived by consumers. The main
 191 aspects of the methodology are summarised in **Table 1**.

192 **Table 1.** Main features of the Water-Energy-Food nexus index methodology.

Features Indicators	Environmental assessment			Nutritional assessment
	Carbon footprint	Water footprint	Energy footprint	Nutritional footprint
Goal and scope	Both environmental and nutritional assessments of any activity related to seafood products for human consumption from fisheries, aquaculture or processing sectors			
Functional unit	1 kg of seafood, either landed at port or produced at the aquaculture or factory gates, including the associated packaging material for processed seafood products			Percentage basis (i.e., 100 g of final product)
System boundaries	Cradle-to- (port/farm/factory) gate			Not applicable
Allocation rules	Mass allocation			Not applicable
Life cycle inventory analysis	To obtain good quality primary data, surveys are recommended to be completed by the responsible agents for further analysis			Nutritional databases
Impact categories	Global warming	Water use (freshwater consumption) and freshwater eutrophication and marine eutrophication (water degradation)	Sum of non-renewable energy (fossil, nuclear and biomass) and renewable energy (biomass, wind, solar, geothermal and water)	Nutrient Rich Food (NRF) index (Drewnowski, 2009)
Characterisation	Global Warming Potentials in a 100-year time horizon (IPCC, 2021)	AWARE (Boulay et al., 2018) for water use and ReCiPe (Struijs et al., 2009) for freshwater eutrophication and marine eutrophication	Cumulative Energy Demand (using Lower Heating Values) (Frischknecht et al., 2007)	NRF12.2, balance between 12 nutrients to be promoted (protein, omega-3, K, Ca, Fe, Mg, I, Se and Vitamins A, C, D and E) and 2 nutrients to be limited (saturated fat and Na)
Units	kg CO ₂ eq.	μPt (after normalisation and weighting of the three impact categories)	MJ	Dimensionless
Normalisation	Reverse linear normalisation of a sample from 0 (highest value of the footprint) to 1 (lowest value of the footprint) (see Equation 5 in Entrena-Barbero et al. (2023a))			Linear normalisation of a sample from 0 (lowest value of the footprint) to 1 (highest value of the footprint) (see Equation 6 in Entrena-Barbero et al. (2023a))
Weighting factors	25 %	25 %	25 %	25 %
Integration	Sum of the four weighted values: Water-Energy-Food nexus index (0 – 100)			

193 The methodology was applied by NEPTUNUS project partners to a series of case studies that can
 194 be divided into three different systems: fisheries, aquaculture and processing. Up to 34 studies
 195 were conducted regarding fisheries, addressing several fishing gears (e.g., trawls, purse seine,

196 pole and line); 7 case studies were assessed with respect to aquaculture; and 22 seafood case
197 studies were related to processing (e.g., freezing, salting, canning and smoking). All case studies
198 were from the European Atlantic area, more precisely in the following countries: France (Cloâtre,
199 2018), Ireland, Portugal and Spain. The timeframe under study was between 2011 and 2019.
200 Given the extraordinary circumstances that occurred in fisheries sector during the COVID-19
201 pandemic in terms of changes in both consumption habits and seafood supply chains, this period
202 was excluded from the sample to avoid misrepresentations in the results obtained (Ruiz-Salmón
203 et al., 2021a). The particularities of each case study can be consulted in more detail in the
204 Supplementary Material (SM), in **Tables SM1-SM3** for the fisheries, aquaculture and processing
205 systems, respectively. Following the WEFni guidelines, the FU was set as 1 kg of seafood, either
206 landed at the port, produced at the farming site, or delivered at the factory gate including the
207 associated packaging material of processed seafood products (Entrena-Barbero et al., 2023b).

208 Once the methodological guidelines were applied to the case studies of seafood products, these
209 were grouped according to fishing gear, type of aquaculture (fish or shellfish) and processing
210 (e.g., salted, smoked, etc.) for the fisheries, aquaculture and processing systems with the purpose
211 of identifying trends. In addition, an analysis of the results obtained was performed for both the
212 single indicators (CF, WF, EF and NF) and the composite index (WEFni). Finally, the strengths
213 and weaknesses of the WEFni methodological guide were discussed.

214 2.2. Sensitivity analysis

215 With the purpose of carrying out a sensitivity analysis of the results obtained, the methodological
216 choices for the estimation of the WEFni, based in Entrena-Barbero et al. (2023a), were modified
217 for two key factors in LCA studies: on the one hand, the FU and, on the other hand, the weighting
218 process as detailed in **Section 2.2.1** and **Section 2.2.2**, respectively. These modifications were
219 applied both individually and jointly, thus obtaining three different scenarios, consisting in the
220 modification of: (i) FU, (ii) weighting process and (iii) both FU and weighting process.

221 2.2.1. Functional unit

222 According to the WEFni guide, a mass-based FU was selected for the assessment of seafood
223 products due to three main reasons: (i) to avoid the common regional differences in price over
224 time (Costello et al., 2020); (ii) it has been the most widely used FU in the seafood LCA literature
225 during the last decade (Ruiz-Salmón et al., 2021b), which facilitates comparisons with other
226 studies; and (iii) it allows for easy and accessible calculations, promoting reproducibility.

227 Despite the foregoing, a portion of mass, in this case 1 kg, does not fully represent the nourish
228 characteristics of food (Weidema and Stylianou, 2020). Thus, to improve the results, it turns out
229 to be essential to include nutritional aspects in the FU of the food production system under
230 evaluation (Heller et al., 2013). Based on this premise, most LCA studies follow three approaches
231 based on: (i) single nutrients, such as the content of digestible proteins (Sonesson et al., 2017),
232 (ii) multiple nutrients in the form of composite indicators (e.g., the NRF), or (iii) diet contexts
233 (McAuliffe et al., 2020).

234 Notwithstanding, during the calculation of the WEFni, the nutritional profile was covered by
235 integrating a specific NF for seafood products (i.e., NRF12.2). Therefore, in this case it was
236 decided to opt for estimating the satiating capacity of seafood products through a caloric- and
237 edible-based FU, named as FU_{ce}. To do this, **Equation 1** was applied, considering the values of
238 the caloric intake (CI) of each specimen. Moreover, since it is not plausible to make comparisons

239 between species with very different edible portions (EPs), e.g., hake and mussel, this factor was
240 also taken into consideration.

241 Likewise, the FU_{ce} was relativised to the amount of seafood in the form of the calories needed to
242 meet the daily per capita demand according to the reference seafood intake (RSI) suggested in the
243 planetary health diet proposed by the EAT-Lancet commission: 40 kcal (Willett et al., 2019). The
244 values taken into consideration and the results obtained for each case study can be consulted in
245 **Table SM4**. On the one hand, case studies having more than one species as target were
246 disregarded, since conducting averages of both CI and EP would lead to obtain an
247 unrepresentative FU_{ce} of the case study. In addition, for processed seafood products containing
248 additives or sauces (e.g., olive oil), as these were already assessed through the NF, only the CIs
249 and EPs of seafood species were considered.

$$FU_{ce} = \frac{RSI}{CI \cdot EP} \quad (1)$$

250 Where:

251 FU_{ce} caloric- and edible-based functional unit (g seafood·(person·day)⁻¹)

252 RSI reference seafood intake per day: 40 kcal·(person·day)⁻¹ (Willett et al., 2019)

253 CI caloric intake of the specimen (kcal·g edible⁻¹)

254 EP edible portion of the specimen (g edible·g seafood⁻¹)

255 2.2.2. *Weighting process*

256 Considering the methodological guidelines of WEFni, the default weighting process carried out
257 was proposed by Andreas et al. (2020), performing a combination of two weighting techniques
258 usually considered in LCA studies. On one hand, panel weighting, which is conducted under the
259 opinion of a group of people, which in this case was the NEPTUNUS project consortium. Partners
260 of the NEPTUNUS project considered to perform a sensitivity analysis giving equal importance
261 to the protection of the environment and to the promotion of nutrition: a weighting factor of 50%
262 for the environmental footprints (with equal weight of 16.67% for CF, WF and EF), as well as a
263 weighting factor of 50% for NF. On the other hand, binary weighting, which considers null or
264 equal weights (i.e., weighting factors of 25% for CF, WF, EF and NF). This was decided under
265 the premise of facilitating the aggregation into a single value (i.e., WEFni) to promote its
266 application and acceptance by producers and consumers, respectively. In addition, none of the 4
267 individual indicators selected were prioritized to show a first simple approximation of what it
268 would mean to follow a benchmarking process based on the WEF nexus concept.

269 3. Results and discussion

270 The results obtained from the case studies after implementing the methodological guidelines are
271 presented according to the values of the single indicators: CF, WF, EF and NF (**Section 3.1**), as
272 well as for the composite indicator: the WEFni (**Section 3.2**). Later, the sensitivity analysis is
273 conducted in **Section 3.3**, to conclude with a discussion of strengths and weaknesses of the
274 approached methodology (**Section 3.4**).

275 3.1. Individual footprints

276 With the aim of conducting the interpretation and discussion of the four individual footprints
277 comprising the WEFni (i.e., CF, WF, WF, and NF), the results are presented in a normalised form

278 (from 0 to 1) in **Figure 2**. The normalisation process was carried out using **Equations 5** and **6**
279 shown in the study by Entrena-Barbero et al. (2023a), taking the maximum and minimum values
280 of the results of each footprint of the whole sample as a reference. Thus, while the seafood product
281 with the lowest environmental footprint in terms of CF, WF or EF was assigned a score of 1, the
282 other seafood products were decreasing their scores proportionally, considering the cases with the
283 maximum environmental footprints with a normalised value of 0. Conversely, since the NF should
284 be as high as possible, the seafood product with the highest and lowest values will become scores
285 of 1 and 0, respectively. In addition, these values have been grouped according to the averages of
286 fishing gears (**Section 3.1.1**), type of aquaculture (**Section 3.1.2**) or processing systems (**Section**
287 **3.1.3**), respectively. For a more detailed information, the values obtained for the four individual
288 indicators in each case study are shown in **Tables SM5-SM7**.

289 3.1.1. Fishing systems

290 In view of **Figure 2A**, up to eight fishing gears integrate the 34 fisheries case studies analysed,
291 being the most numerous purse seine, trawls, and artisanal fisheries with 12, 8 and 7 cases,
292 respectively, while others only included 1 or 2 case studies (pole and line, pots and traps, dredge,
293 trammel net and longline).

294 In terms of environmental footprints (CF, WF and EF), the top three fishing gears with the lowest
295 environmental burdens (and thus had the largest individual normalised scores) were purse seine,
296 artisanal fisheries, and dredge. In contrast, the highest environmental burdens (i.e., the lowest
297 normalised scores) were reported for longline and pots and traps. Regarding the CF, the average
298 values without normalisation ranged between 0.56 and 3.33 kg CO₂ eq. per kg of landed fish for
299 purse seiners and longlines respectively. Between these two limit values, the following fishing
300 gears are ranked in order of highest to lowest impact: pots and traps (with 3.08 kg CO₂ eq.),
301 trammel net and trawls (with similar values of 2.17 and 2.01 kg CO₂ eq., respectively) and dredge
302 and artisanal fisheries, which reached a similar position at around 0.80 kg CO₂ eq. on average. It
303 should be noted that the average trawl aggregation includes both pelagic and bottom trawlers,
304 with CF values ranging from 0.38 kg CO₂ eq. for the former, to 3.31 kg CO₂ eq. for the latter.
305 Further analysis of the factors contributing to the total CF indicated that the major hotspot was,
306 in most cases, the use of fuel during fishing.

307 To contrast the results, the most representative samples — trawl and purse seine case studies —
308 were compared with the existing literature. A recent study by Sandison et al. (2021) examined the
309 environmental performance of Scottish pelagic fisheries by applying the LCA methodology with
310 a similar scope and using 1 kg of whole mixed pelagic fish as the FU. The sample consisted of
311 eleven trawlers, three of which combined trawl and purse seine gears. The authors reported an
312 average CF of 0.45 kg CO₂ eq. per FU. However, they analysed inter-vessel and inter-annual
313 comparisons and observed CF fluctuations ranging from a minimum of 0.28 kg CO₂ eq. to a
314 maximum of 0.74 kg CO₂ eq., and a value of 0.23 kg CO₂ eq. to 0.29 kg CO₂ eq. for each scenario,
315 respectively. The results found in this study for pelagic trawlers were in line with the Scottish
316 fleet average (Sandison et al. 2021). On the other hand, the average CF obtained for purse seine
317 (0.56 kg CO₂ eq.) is in line with the value reported in a study of the Galician (Spain) purse seine
318 fleet (30 vessels), which accounted 0.79 kg CO₂ eq. (Vázquez-Rowe et al., 2010).

319 On average terms, WF results showed a similar trend to the CF, moving in a range from 7.02 to
320 37.45 μPt per kg of landed fish, being purse seine and longline the fishing gears with the lowest
321 and highest values, respectively. A recent study by Ceballos-Santos et al. (2023) obtained a WF
322 of 11.98 μPt for the Cantabrian purse seine fleet here evaluated. This value was the aggregation

323 of the individual indicators (water use: $6.23 \cdot 10^{-2} \text{m}^3 \text{eq.}$, freshwater eutrophication: $7.38 \cdot 10^{-6} \text{kg P}$
324 eq. , and marine eutrophication: $7.93 \cdot 10^{-3} \text{kg N eq.}$). Ceballos-Santos et al. (2023) also analysed
325 the environmental impacts in depth, revealing that the main contributor to the WF was diesel
326 production and combustion, which accounted for 99% of the overall marine eutrophication
327 indicator (this marker contributed 95% to the clustering of water footprint indicators). The same
328 occurred for the longline, in which the marine diesel necessary for the fishing of sardines used as
329 bait entailed practically all burdens in WF terms. Likewise, the longlining value was followed by
330 trammel nets (30.53 μPt) and trawls (28.85 μPt), whereas pots and traps showed an intermediate
331 footprint of 21.6 μPt . On the other side, artisanal fisheries and dredges presented similar average
332 values of 10.85 and 10.81 μPt , respectively. These values come from the normalisation and
333 weighting of the results of three impact categories related to both freshwater consumption (water
334 use) and water degradation (freshwater eutrophication and marine eutrophication) hindering the
335 comparison with reference values reported in literature.

336 With regard to the non-normalised EF results, they were between 8.45 MJ for purse seine and
337 49.78 MJ for pots and traps, per kg of landed fish, following the pattern observed for CF and WF.
338 The values of the latter were followed closely by longline (41.17 MJ), while trammel net recorded
339 more distant results (33.4 MJ). Behind of these results were the obtained by trawls (30.02 MJ)
340 and pole and line (26.15 MJ), while dredge and artisanal fisheries showed EFs of 11.54 MJ and
341 9.79 MJ, respectively. Once again, it was noted that the energy required for the production and
342 combustion of diesel was responsible for most environmental loads encompassed in the EF.

343 Against this backdrop it is interesting to examine how footprints vary according to how seafood
344 is produced, i.e., the same species captured using different fishing gears. For instance, Atlantic
345 mackerel (*Scomber scombrus*) is the target species of French and Galician trawlers, Cantabrian
346 purse seiners, as well as artisanal fisheries fleets. In such case, the highest footprint values were
347 obtained by Galician trawlers, with a CF of 1.33 kg CO₂ eq., a WF of 16.17 μPt and an EF of
348 17.83 MJ per kg of Atlantic mackerel captured. For the case of French trawlers, despite being the
349 same kind of gear, the footprint values were around half: 0.52 kg CO₂ eq., 6.93 μPt and 7.79 MJ
350 per FU. Regarding Cantabrian purse seiners, the lowest footprint values were registered: 0.11 kg
351 CO₂ eq., 0.34 μPt and 1.59 MJ per kg of Atlantic mackerel captured, while the use of artisanal
352 fisheries presented intermediate values between those obtained by trawlers and seiners: 0.82 kg
353 CO₂ eq., 13.16 μPt and 1.73 MJ. Another example is the fishing of albacore tuna (*Thunnus*
354 *alalunga*) by French trawlers, Cantabrian seiners and artisanal fisheries. For this case, the trawling
355 gear also presented the highest footprint values, with a CF value of 4.25 kg CO₂ eq., 58.18 μPt
356 and 65.36 MJ per kg of albacore tuna captured. In contrast, artisanal fisheries fleet had the lowest
357 CF and EF values: 0.13 kg CO₂ eq. and 0.26 MJ, respectively, while purse seiners obtained the
358 lowest WF: 0.74 μPt . These results are in line with the well-documented literature that points out
359 that purse seine fleets targeting inshore fish stand out as one of the fishing gears that generate the
360 least environmental burdens since they are coastal and low fuel-consuming fisheries (Vázquez-
361 Rowe et al., 2010). In contrast, several studies around the world have found that trawlers are the
362 fleet group with the highest fuel consumption per kg of fish, partly because of the long distance
363 they travel to the fishing grounds (Jafarzadeh et al., 2016; Sandison et al., 2021). In fact, bottom
364 trawling was found to be up to five times less efficient than purse seining in some cases (Hognes
365 et al., 2012; Schau et al., 2009).

366 Regarding the NF scores, it is important to take into consideration that the nutritional assessment
367 depends only on the species of seafood captured and not on the fishing gear used. Consequently,
368 for the fishery scenario, around 25 species were assessed using the dimensionless NRF12.2 index,

369 involving from top predators as tuna (e.g., *Thunnus alalunga*) to small pelagic as sardine (*Sardina*
370 *pilchardus*), along with cephalopods (e.g., *Octopus vulgaris* and *Sepiida*) and bivalves (e.g.,
371 *Pecten maximus*). The NF values ranged between 115 and 361, being ~~Chub-chub~~ mackerel
372 (*Scomber japonicus*) and the mix of sea bass (*Dicentrarchus labrax*) and gilthead seabream
373 (*Sparus aurata*) the species with the highest and lowest NF values, respectively. No large
374 contrasts were identified between pelagic and demersal species which reached an average
375 NRF12.2 index of 234 and 213, respectively. Cephalopods accomplished the highest mean score
376 (271) and bivalves the lowest with 187 points. The low NRF12.2 value for the latter is due to the
377 absence of omega-3 content, one of the most valuable nutrients to be promoted in seafood,
378 coupled with a high amount of saturated fat (0.23 g per 100 g), one of the nutrients to be limited
379 in the applied model. These outcomes are quite different from Koehn et al. (2022), who used a
380 modified version of the nutrient richness index developed by Drewnowski (2009), demonstrating
381 that small pelagic fish rated highest than bivalves and cephalopods. In this line, Bianchi et al.
382 (2022) found large variability within species groups in terms of nutrient density scores, but
383 suggested that wild salmonids and small pelagic performed the best.

384 3.1.2. Aquaculture systems

385 The results from aquaculture systems (**Figure 2B**) were divided into four fish and three shellfish
386 case studies. Globally, shellfish aquaculture achieved higher normalised scores (almost double)
387 for all individual footprints, meaning lower environmental impacts than farmed fish. In terms of
388 CF, the results per FU ranged from 0.08 for oysters (*Crassostrea gigas* and *Crassostrea angulata*)
389 to 10.30 kg CO₂ eq. for gilthead seabream (*Sparus aurata*), with ten times higher average values
390 for shellfish (0.55 kg CO₂ eq.) compared to fish aquaculture (5.21 kg CO₂ eq.). For the WF this
391 deviation increased, with averages values of 2.16 µPt for shellfish and 54.66 µPt for fish species.
392 As for the EF, the same differences recorded in the CF occurred, with average values for shellfish
393 and fish of 10.00 and 104.35 MJ per FU, respectively, being the maximum for the farming of
394 gilthead seabream (211 MJ). These wide ranges in terms of minimum and maximum values for
395 the environmental indicators could be related to the type of aquaculture implemented in the case
396 studies analysed. For example, those based on intensive aquaculture need a more controlled
397 production process in terms of environmental conditions, external feed or existence of an artificial
398 flow of water that need to be pumped, which translates into higher CF, WF and EF. Therefore,
399 given the particularity of each case study depending on the location and species assessed, it is
400 difficult to make equivalent comparisons. Finally, the nutritional quality of aquaculture products
401 calculated through the NF showed punctuations from 194 for gilthead seabream to 314 for
402 rainbow trout (*Oncorhynchus mykiss*), with shellfish and fish species obtaining a similar average
403 NF of 229.04 and 232.49, respectively. These findings were consistent with those of Gephart and
404 Golden (2022). These authors ranked a sample of aquaculture groups according to the total
405 environmental and nutrient scores, showing that bivalves performed best, followed by silver and
406 bighead carp, trout, and salmon. However, they notified significant heterogeneity in both
407 environmental performance and nutrient content performance, with almost all blue food groups
408 scoring best in at least one metric, leading to possible trade-offs in the assessment of the double
409 bottom line in aquaculture.

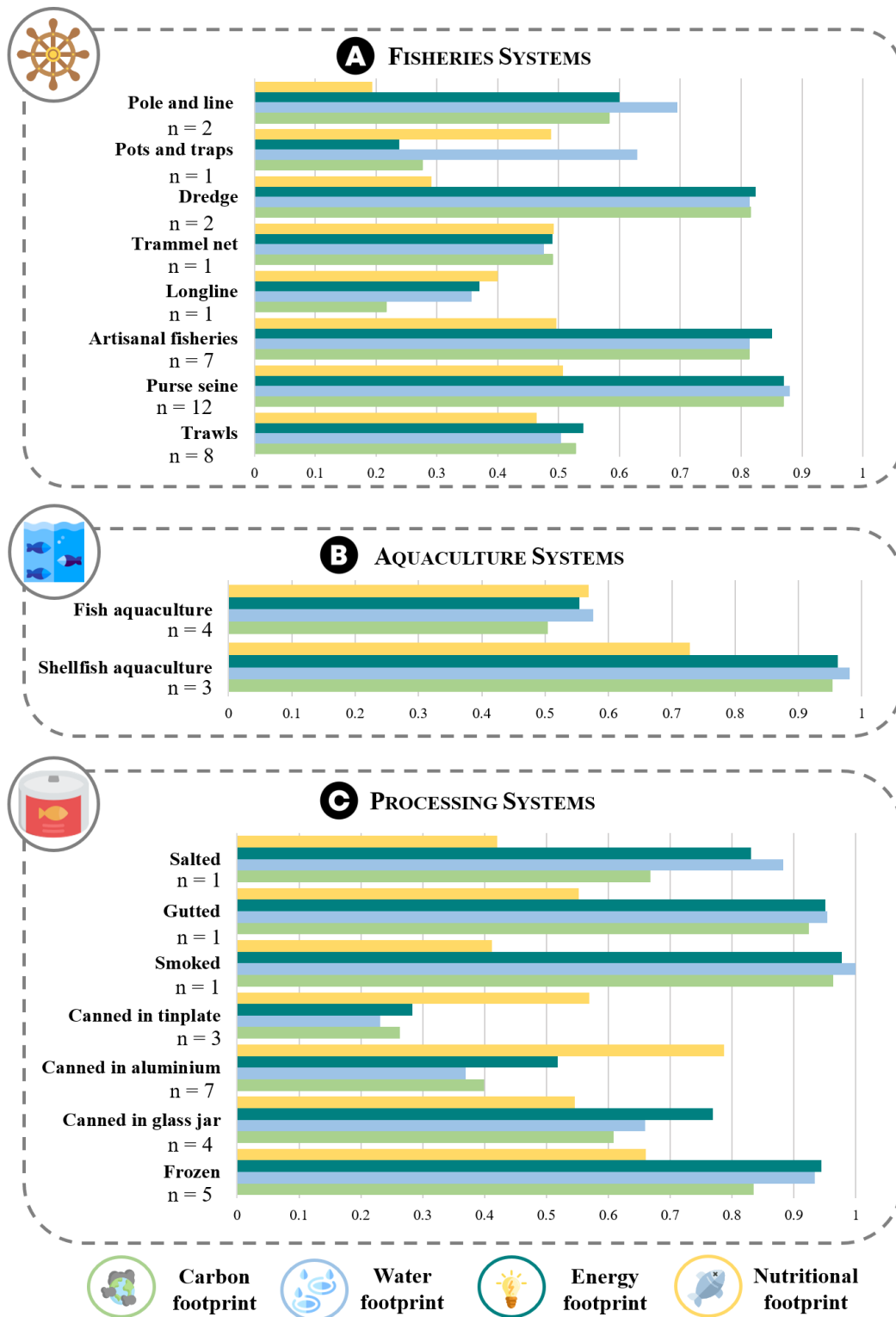
410 3.1.3. Processing systems

411 With the focus on the 22 case studies of processed products analysed (**Figure 2C**), these were
412 grouped into different processing techniques: freezing (5 case studies); canning in glass jar (4),
413 aluminium (7) and tinplate (3); and smoking, gutting, and salting, with one case study for each

414 type. In addition, different scenarios related to the incorporation of diverse oils and other additives
415 were assessed for canning processes (see **Tables SM1-SM3** for further information). In view of
416 **Figure 2C**, smoked, gutted and frozen products highlighted for reaching upper normalised scores,
417 thus related to having lower footprints. The smoked product got the lowest average CF per FU,
418 with 1.10 kg CO₂ eq., followed by the gutted product (1.61 kg CO₂ eq.) and frozen products (2.77
419 kg CO₂ eq.). The same ranking was observed for WF, with results of 7.30, 23.69 and 30.69 μPt,
420 for smoked, gutted and frozen items, respectively. In contrast, the gutted product achieved an EF
421 score (18.61 MJ) ten points lower than smoking (28.55 MJ). On the opposite side, canned goods
422 showed the worst performance, with tinplate canned products showing higher averages in CF per
423 FU (10.19 kg CO₂ eq.), WF (278.54 μPt) and EF (282.92 MJ). Then, products canned in
424 aluminium and glass jars followed behind tinplate case studies with CFs of 8.42 and 5.71 kg CO₂
425 eq., respectively. In terms of WF 229.67 and 127.33 μPt were obtained for products canned in
426 aluminium and glass jar, respectively. As for the EF, canning in aluminium reached an impact
427 (193.51 MJ) almost two times above canning in glass jars (98.13 MJ). Finally, the average salting
428 results were intermediate (CF: 4.94 kg CO₂ eq.; WF: 48.75 μPt; EF: 74.41 MJ). Comparing the
429 most abundant type of processing in the sample, canning, with existing studies it was found that
430 the average CF was in the range of values reported for mussels canned in tinplate, with 4.32 kg
431 CO₂ eq. per kg. (Iribarren et al., 2010) and tuna canned in tinplate, with 8.4 kg CO₂ eq. per kg
432 (Avadí et al., 2015). If a more in-depth analysis is carried out, the stage of seafood production
433 (i.e., fishing or farming) usually has a significant contribution to the environmental burdens.
434 Therefore, if for example a comparison is carried out between frozen cod and sardine, results can
435 vary greatly because

436 On the nutrition side, NF values for all case studies were within the range of 101-306 with smoked
437 and canned in aluminium products having the lowest and highest averages (185.40 and 262.14,
438 respectively). The nutritional difference between case studies with the same fish species was due
439 to the rest of ingredients used and their proportion to elaborate the final product.

440 In summary, after analysing the individual footprints of each group per sector, it can be concluded
441 that, although some trends have been identified, it would be necessary to include a larger and
442 more diverse sample in future iterations to enable cross-system comparisons to obtain more
443 plausible results for the Atlantic area. On the other hand, a more detailed study of hotspots has
444 not been carried out due to the lack of equality in the resolution of results for the whole sample
445 and confidentiality issues related to the inventories. Limitations in terms of availability and
446 representativeness of the data used in each case study could have changed the results to a
447 significant extent. Likewise, it should not be overlooked that the main objective of the study is
448 not to identify hotspots of each case study, but rather to systematically quantify the impacts of
449 case studies of marine systems in the European Atlantic area.



450

451 **Figure 2.** Individual indicators (i.e., carbon, water, energy and nutritional footprints) for each system
 452 assessed: (A) fisheries, (B) aquaculture and (C) processing. The results are presented normalised and
 453 averaged for the different groups within each system. The lowest value in terms of carbon, water and energy
 454 footprint is assigned a score of 1, while the remaining products decrease in proportion, considering 0 as
 455 the highest footprint. Conversely, for the nutritional footprint the higher the better, so the highest and lowest
 456 values are assigned with scores of 1 and 0, respectively.

457

458 3.2. Composite indicator

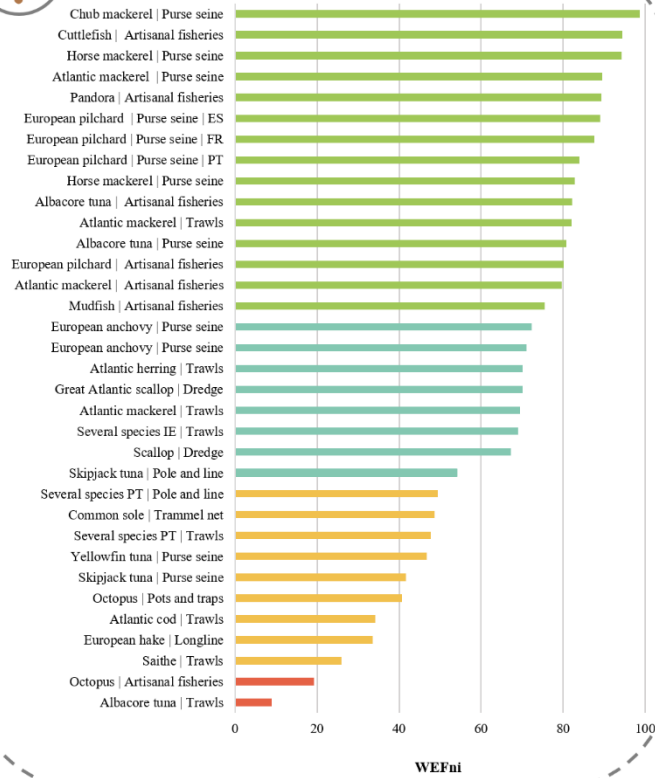
459 The four individual footprints shown above (i.e., CF, WF, EF, NF) were aggregated into a single
460 value (i.e., the WEFni), as shown in **Figure 3**, for each case study, according to the type of system:
461 fisheries (**Figure 3A**), aquaculture (**Figure 3B**), and processing (**Figure 3C**).

462 For the fisheries systems (**Figure 3A**), the list was mainly headed by species captured using purse
463 seine and artisanal fisheries. At the top was the case of chub mackerel purse seine with 99 out of
464 100 points. Behind this case were cuttlefish captured through artisanal fisheries and horse
465 mackerel captured through purse seiners, being scored with a WEFni of 94 points. A further 12
466 case studies exceeded 75 points (thus being green coloured), including not only different species
467 landed with the abovementioned fishing gears, but also the trawl fishery for Atlantic mackerel
468 (82 points). Shown as blue bars (varying the WEFni from 51 to 75 points), up to seven case studies
469 have an average value of approximately 70, being mainly represented by purse seiners and
470 trawlers, while pole and line catches of tuna are below 60. Subsequently, with scores between 26
471 and 50 points (yellow coloured) were 12 case studies comprising different types of fishing gears
472 (trammel net, longline, pots and traps, etc.) and species, such as common sole, saithe or octopus.
473 Finally, at the bottom (25 or less points and red coloured) there were two cases: octopus captured
474 with artisanal fisheries (19 points), and the albacore trawl fishery (9 points). When a detailed
475 comparison of fishing gears was carried out by setting mean WEFni values, purse seiners scored
476 the highest (69 points), followed by artisanal fisheries (66 points), dredges (64 points) and
477 trammel nets (56 points). Pole and line, trawls, and pots and traps obtained lower values, with 48,
478 45 and 32 points, respectively, while longline showed the worst WEFni performance (27 points).

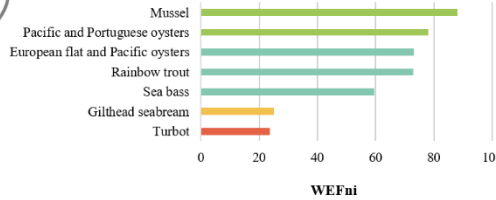
479 When it came to the aquaculture products (**Figure 3B**), shellfish were found to score best, with
480 mussels reaching 88 points, followed by Pacific and Portuguese oysters, which scored 78 points.
481 In fact, this oyster mix performed slightly better than the one composed of European flat and
482 Pacific oysters, which reached 73 points, having a similar performance to farming of rainbow
483 trout and sea bass. At the end, with the lowest points, 25 and 24, were gilthead seabream and
484 turbot, respectively. The results for the WEFni showed the same trend as those found from the
485 individual footprints. Even though normalisation among case studies, it was difficult to draw clear
486 conclusions, despite some trends emerged showing a better performance of bivalves when
487 compared to fish farming.



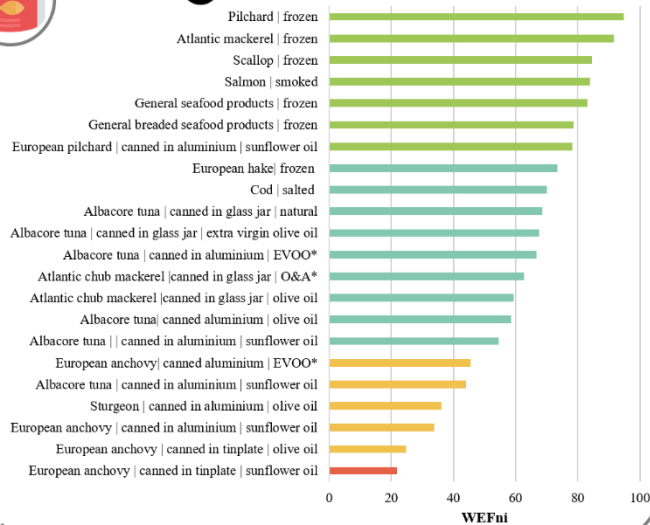
A FISHERIES SYSTEMS



B AQUACULTURE SYSTEMS



C PROCESSING SYSTEMS



490 **Figure 3.** Water-Energy-Food nexus index (WEFni) results for each case study aggregated by system type:
 491 (A) fisheries, (B) aquaculture, and (C) processing. *For fisheries systems: this purse seine case study for
 492 the European anchovy is landed in a non-European port. For processing systems: EVOO denotes Extra
 493 Virgin Olive Oil and O&A denotes Olives and Almonds as additive.

494 With regard to case studies of activities belonging to the seafood processing industry (**Figure**
 495 **3C**), 5 frozen products were found with a score above 76 points, with pilchard freezing showing
 496 the best performance (95 points). The other case studies concerned species ranging from salmon
 497 or scallops to breaded products and frozen seafood in general. The smoked salmon scenario also
 498 reached a high score (84 points). In addition, a canning scenario entered this top position, with 78
 499 points for European pilchard with sunflower oil canned in aluminium. In the intermediate values,
 500 between 26 and 75 scores, most canned products were scored, covering a variety of species,
 501 additives and packaging formats. In addition, the cases of frozen European hake (73) and salted
 502 cod (70) were also scored in intermediate levels. Among the different canning scenarios, albacore
 503 tuna and Atlantic chub mackerel in glass jars achieved better scores, with natural products (i.e.,
 504 without additives) followed by those with extra virgin olive oil (less refined than olive oil or
 505 sunflower oil) scoring the highest WEFni. On the other hand, European anchovy canned in
 506 tinplate with olive oil scored the worst with only 22 marks. These outcomes may suggest that the
 507 use of glass jar packaging for canned products results in a lower environmental burden when
 508 compared with aluminium and tinplate cans, thus suggesting that the type of packaging material
 509 can improve the environmental performance of these preserved products (Almeida et al., 2022).
 510 The ranking of canned scenarios revealed clearer conclusions using the composite indicator
 511 compared with the individual footprints.

512 3.3. Sensitivity analysis

513 The WEFni obtained for the different case studies after carrying out the sensitivity analysis
 514 applied to the FU, the weighting process and the consideration of both approaches are compiled
 515 in **Figure 4**, classified in fisheries (A), aquaculture (B), or processing (C) systems. Moreover,
 516 given the large number of case studies to which sensitivity analysis was applied (58), it was
 517 decided to group them in a similar way as shown in **Figure 2**. Notwithstanding, the individual
 518 results for each case study have been compiled in **Tables SM8-SM10**.

519 Regarding the fishing systems (**Figure 4A**), when the sensitivity analysis was conducted in
 520 relation to the FU, it was observed that all fishing gears, except trawls and purse seines, decreased
 521 their WEFni. This was triggered by the FU_{ce} assigned to each target species. A lower FU_{ce}
 522 translates into a lower amount of seafood (in mass) to satisfy the RSI per person per day: 40 kcal
 523 (Willett et al., 2019), resulting in a lower environmental impact (i.e., lower CF, WF, and EF) and
 524 a higher WEFni. Moreover, the FU_{ce} assigned to each species depends in turn on both its caloric
 525 intake and its edible portion (i.e., CC and EP, respectively). Thus, it was identified that for several
 526 species captured by trawls or purse seines (e.g., saithe, European pilchard or Atlantic mackerel),
 527 the lowest FU_{ce} were recorded, varying approximately between 35 and 45 g of catches per person
 528 per day (see **Table SM4**). Conversely, more than 90 g of catches were required per person per
 529 day to satisfy the RSI through pots and traps (octopus) and trammel net (common sole) thus
 530 leading to the largest reductions in terms of WEFni.

531 Similar results were reported when the weighting process was modified. In this case, the NF
 532 associated with each species turns out to be now the key factor in increasing or decreasing the

533 WEFni. This was because, by giving equal importance to the environment and food pillars, the
534 weighting factors assigned to each environmental footprint (i.e., CF, WF, and EF) has decreased
535 with respect to the baseline situation (from 25% to 16.67%), while the nutritional aspects have
536 seen their contribution double: from 25% to 50%. Therefore, the increase in the WEFni for trawls
537 and purse seines can be justified from the point of view that the target species of these fishing
538 gears were the ones with the highest NFs (see **Table SM5**), showing values usually higher than
539 270 for the NRF12.2 (e.g., Atlantic cod, chub mackerel and horse mackerel). Consequently, this
540 positioned them above the average of the whole sample of fisheries systems: around 230. On the
541 opposite side were the low NF of European hake (213.61) and skipjack tuna (210.32), which
542 implied that longline (minus 54 points) and pole and line (minus 50 points) revealed the greatest
543 reductions in terms of WEFni.

544 Likewise, when both approaches (FU and weighting process) are considered within the sensitivity
545 analysis of fisheries systems, a similar trend was observed once again, being trawling and purse
546 seine the only groups which increased its WEFni by 31 and 8 points, respectively. Conversely,
547 pots and traps, dredge, longline and trammel net were the most disadvantaged groups with
548 reductions from 54 to 42 points. Finally, when all fishing gears are compared with each other,
549 artisanal fisheries and purse seine benefit the most from the sensitivity analysis, moving from
550 sixth and seventh place respectively to at least the top two according to the WEFni. In the opposite
551 situation was pots and traps, which has slipped from second place to the bottom of the ranking.

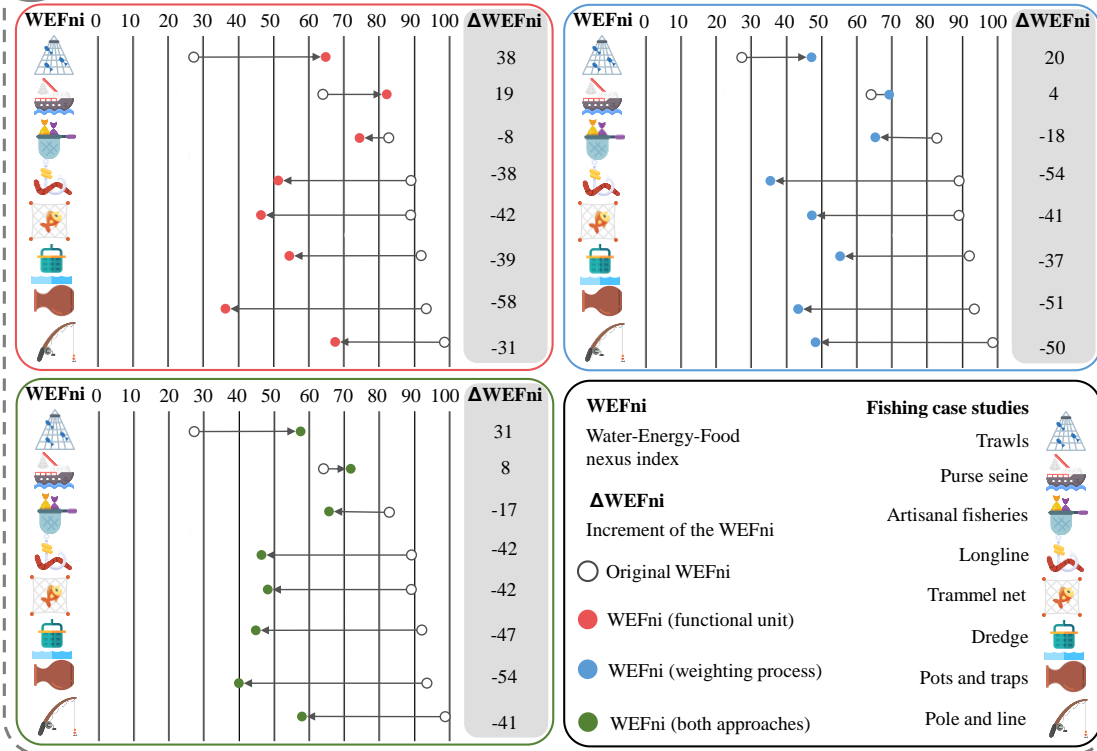
552 Concerning aquaculture systems (**Figure 4B**), after conducting the sensitivity analysis for the two
553 groups, fish and shellfish aquaculture, few differences were identified. The FU_{ce} of the fish
554 aquaculture group has on average 81.49 g of catches per person per day to meet the RSI, a much
555 lower value compared to shellfish aquaculture (266.67 g and 416.67 g for mussels and oysters,
556 respectively). This translates into fish aquaculture improving its WEFni by 3 points, while
557 shellfish aquaculture decreased 8 points. Moreover, the sensitivity analysis compromising the
558 weighting process implies reductions in both groups, with a higher decrease in shellfish than in
559 fish. This was because fish have a worse environmental performance in terms of CF, WF, and EF;
560 thus, in principle, this should have increased their WEFni by decreasing the total weighting of
561 these three footprints (from 75% to 50%). However, by increasing the weighting factor of the NF
562 the WEFni was decreased, as certain fish register worse values, such as gilthead seabream
563 (194.06) or turbot (207.02) compared to the shellfish species analysed, i.e., oyster and mussel,
564 with NF equal to 210.00 and 267.12, respectively. When the two approaches were addressed
565 together, it became clear that fish aquaculture was catching up shellfish aquaculture, with 15
566 points separating them in terms of WEFni. Lastly, according to the ranking, although both groups
567 tend to reduce their differences no change in positions occurred, with fish aquaculture remaining
568 in first place and shellfish aquaculture in second.

569 In relation to processing systems (**Figure 4C**), the sensitivity analysis of the FU affected mostly
570 salted and gutted products, decreasing their WEFni by 11 and 22 points, respectively, while for
571 the remaining processed systems, i.e., smoked, canned and frozen, the results remained almost
572 unchanged. This can be explained by the FU_{ce} reported by each case study for the salted and gutted
573 processes: 72.07 and 190.48 g of catches per person per day for Atlantic cod and Atlantic scallop,
574 respectively. These values are affected by the low EP of the great Atlantic scallop (25%) and by
575 the low CC of cod (74 kcal per 100 g edible). Consequently, both case studies were in
576 disadvantage compared to the others, being mainly composed by species such as albacore tuna,
577 European anchovy or Atlantic chub mackerel, which contain high values for both CC and EP,

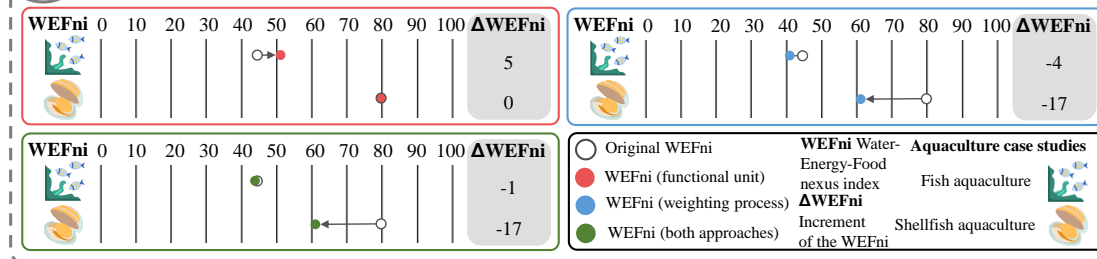
578 thus having a much lower FUce (46.71 g of catches per person per day). The modification of the
579 weighting process led to small reductions in the WEFni for canning (1 point) and freezing (5
580 points) process systems. Conversely, for salting, gutted and smoking processes, this variation
581 implied a reduction of at least 10 points, caused by the low NFs of cod, great Atlantic scallop and
582 salmon, respectively. These trends were exacerbated when the combined effects of FU and
583 weighting process were analysed, with further reductions in the WEFni for salting (17 points) and
584 gutted (24 points) processing systems. There were little variations in the ranking of the different
585 types of processing, with frozen products still occupying the first place, while canned and salted
586 products were the worst WEFni performers.



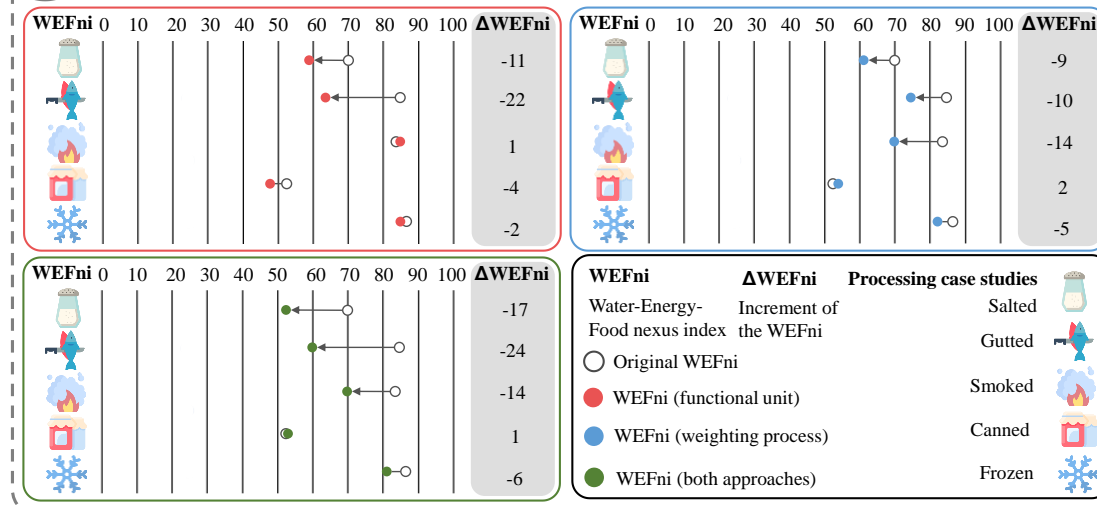
A FISHERIES SYSTEMS



B AQUACULTURE SYSTEMS



C PROCESSING SYSTEMS



588 **Figure 4.** Water-Energy-Food nexus index results of the sensitivity analysis for each case study divided by
589 system: (A) fisheries, (B) aquaculture, and (C) processing. White dots represent the original WEFni
590 obtained in the base case; red line and dots show the results of the sensitivity analysis performed by varying
591 the FU; line and dots in blue show the results of the sensitivity analysis carried out varying the weighting
592 factors; and line and dots in green show the results for the sensitivity analysis carried out with both
593 approaches jointly.

594 3.4. Strengths and weaknesses of the WEFni methodological guide

595 The methodological guidelines applied in this work allowed to assess the environmental impacts
596 and nutritional profiles of a sample of seafood species from the European Atlantic area through a
597 composite index: the WEFni. In addition, this information can be represented as a percentage in
598 the form of a front-end ecolabel to be implemented in seafood products packaging. For that, a
599 design proposal was suggested by Entrena-Barbero et al. (2023a) with the goal to facilitate the
600 communication between producers and consumers. However, while the ecolabel can be a useful
601 tool to consumers that wish to make more sustainable purchases, the representation of results
602 through a single value can mask the main hotspots of the single indicators that compose the
603 WEFni for a more aware or expert audience. In this regard, methodological transparency is key.
604 Further developments in the WEFni ecolabel could include the provision of more detailed
605 information through online channels (e.g., by using a QR code) about the outcomes and
606 methodological procedures of the four individual footprints. In this regard, it is important to bear
607 in mind that the aims and limitations of composite indicators should be presented alongside where
608 the scores come from, what do they mean and what are the limits of their usefulness or
609 interpretability (Barclay et al., 2019).

610 Since the WEFni methodology implemented in the seafood sector can be applied also to different
611 food products, this should be adapted to the particularities of each foodstuff. In this regard, one
612 of the main aspects to be modified would be the "food" pillar within the WEF nexus (i.e., NF).
613 To do this, the NRF12.2 index used for the seafood products could be modified by adding or
614 disregarding certain nutrients to better assess the nutritional profile of products under evaluation.
615 However, it is also possible that certain aspects of the food pillar may be approached from the
616 point of view of ensuring food security in terms of productivity. This perspective was addressed
617 in the case of a WEFni proposal applied to dairy farms, where a milk yield indicator was
618 considered (Entrena-Barbero et al., 2023c). Likewise, it is important to take into consideration
619 that the population is not exclusively nourished by seafood products purchased at points of sale
620 (supermarket, groceries, etc.). Therefore, this WEF nexus thinking should be extended to different
621 food items that are consumed, as well as in other forms of food consumption such as those out-
622 of-home: meals and menus related to the HORECA (Hotels, restaurants and catering) channel.

623 To address the WEF nexus perspective, there are numerous methodologies to represent in both a
624 qualitative and quantitative manner depending on the objective. Among them, concept maps or
625 causal loop diagrams can be considered to represent at abstract level the interconnections within
626 a WEF nexus system, while other modelling can be applied in the case of quantitative results,
627 such as agent-based and input-output modelling (Sušnik et al., 2022). In this paper an LCA was
628 followed, which stands out from other methodologies for the holistic quantification of
629 environmental burdens related to the interrelationships of the nexus by considering the resources
630 consumed throughout the life cycle of a product (Mannan et al., 2018).

631 Notwithstanding, the choice of approaching an LCA has in turn certain limitations related with
632 methodological assumptions. Some are related with the management of by-catch from fisheries.

633 While some case studies considered them as co-products and used mass allocation procedures to
634 assign the corresponding environmental loads to them, others treated these as waste (output)
635 without clarifying their fate, which should be noted as a limitation when making comparisons and
636 interpreting results. In terms of data limitation, gaps related to non-available specific inputs were
637 filled by using the best available generic or extrapolated data from feasible sources. The
638 geographical and temporal coverage is also considered to generate some deviations in the general
639 guidelines, for instance certain hypotheses or modelling assumptions could be considered
640 obsolete. This again signals that the proposed guidelines should be kept up to date. Another key
641 aspect is the FU selected. Being 1 kg of seafood captured does not address the nutritional quality
642 of seafood products (FAO, 2021). The study of how the nutritional characteristics could modify
643 the results obtained has been addressed through a sensitivity analysis in which the CC, as well as
644 EP of each species were integrated under the basis of an RSI.

645 As regards environmental impact assessment, although global warming (CF), water scarcity (WF)
646 and energy requirements (EF) have important interrelations with food security, some other impact
647 categories have been left out of the scope of the study. For the case of seafood products, it is
648 relevant to include the ecological status of fishing grounds and its associated biodiversity loss. In
649 this concern, new LCA impact categories have been proposed as, for example, one that quantifies
650 ecosystem quality for fisheries based on several key aspects such as relative biomass, growth rate
651 and biotic resource depletion (Hélias et al., 2023). In addition, it would also be interesting to
652 include the impact associated to plastic leakage along the supply chain of seafood, that occur
653 during fishing activities (fishing gear loss), transportation (tyre particle loss) and processing and
654 end-of-life (plastic packaging loss) (Loubet et al., 2022). Such integration would enable to
655 discriminate different fishing methods associated with various fishing gear loss, different
656 packaging or end-of-life options for fishing gear and packaging (Schneider et al., 2023).

657 Lastly, to broaden the focus of the WEFni towards a sustainability point of view, as it only
658 considers the environmental pillar, the addition of socio-economic indicators would be important.
659 Examples of the above that apply to the seafood sector include human and labour rights, economic
660 viability, and ethics criteria (e.g., animal welfare). Some are already included in certification
661 schemes such as Friends of the Sea, Naturland Sustainable Capture Fishery of Marine, and
662 Aquaculture Stewardship Council (International Trade Centre, 2023).

663 **4. Conclusions**

664 In the current climate emergency situation, reducing the environmental impacts of food
665 production systems is key to reach climate goals and ensure global food security. Therefore, a
666 good option would be to modify dietary patterns by opting for foods with low environmental
667 burdens and high nutritional profiles, such as blue food. To facilitate this task, many ecolabels
668 exist on the market for seafood products, but most focus on a simple environmental indicator and
669 it is crucial to consider multiple interrelations. For this purpose, the WEFni methodology was
670 applied to a series of case studies in the European Atlantic area. From the results obtained, it can
671 be concluded that the WEFni makes each case study more easily interpretable, as they are
672 influenced by numerous particularities such as fishing gear, target species, location or processing
673 technique. In addition, general trends were found to allow the identification of fisheries,
674 aquaculture or seafood processing practices that can be more recommendable from a WEF nexus
675 perspective.

676 The results of the sensitivity analysis carried out showed that the WEFni was considerably
677 modified, implying a reclassification of the fishing gear groups, as well as for the processing
678 types, while for aquaculture systems, some improvement was identified for fish farming
679 compared to shellfish farming. This therefore suggests that, on the one hand, a re-evaluation of
680 the chosen mass FU may be necessary, as the FU_{cc} has led to a better definition of the true function
681 of seafood production systems: to meet the recommended daily personal demand (i.e., the RSI),
682 ensuring food security within the WEF nexus. On the other hand, to better represent the
683 perspective of the WEF nexus in the context of the study, the process of equitable weighting of
684 the individual indicators could also be reconsidered when obtaining a WEFni.

685 The outcomes provided in this paper may serve to apply the WEFni methodology to other case
686 studies, thus expanding LCA studies on seafood products. Besides, In addition, with appropriate
687 modifications that address the particularities of each sector, the WEFni ecolabel could be
688 extended to other food products and distribution channels.

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