

1 **Developing a site index model for *P. pinaster* stands in NW Spain**  
2 **by combining bi-temporal ALS data and environmental data**

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24 **Abstract:** Site index (SI) is a common measure of forest site productivity, serving as a  
25 valuable baseline for forest management. The main objective of this study was to develop  
26 a SI model for *Pinus pinaster* Ait. in north-west Spain by combining bi-temporal,  
27 low-density airborne laser scanning (ALS) data (acquired in the periods 2009-2011 and  
28 2015-2017) with climatic, edaphic and physiographical data. Site productivity, assessed by  
29 site quality curves, was modelled using an age-independent difference equation method  
30 based on ALS metrics and environmental variables. For the model development process,  
31 we used data from 156 sample plots in pure and even-aged *P. pinaster* stands distributed  
32 throughout Galicia (NW Spain) and measured in the Spanish National Forest Inventory  
33 (SNFI). The generalized algebraic difference approach (GADA) formulation was tested by  
34 using two different base equations for modelling the dominant height growth ( $\Delta H$ ) from  
35 ALS variables. The GADA formulation derived from the Bertalanffy's base model  
36 produced the best estimates of dominant height ( $H$ ) for *P. pinaster* stands in Galicia. Use  
37 of the proposed model to estimate  $\Delta H$  for a new pine stand requires two ALS data sets for  
38 estimating site-specific (local) parameters. To enable use of the model when such  
39 information is not available, the relationship between the values of the site-specific  
40 parameter and environmental variables was described using Multivariate Adaptive  
41 Regression Splines (MARS). Use of the MARS equation enabled us to develop spatially-  
42 explicit predictive maps of the site-specific parameter values, which can be used together  
43 with the GADA model to derive  $\Delta H$  curves and SI estimates for *P. pinaster* stands in the  
44 whole study region.

45

46 **Keywords:** LiDAR; Forest productivity; Forest growth modelling; Multi-temporal data

47

## 48 **Introduction**

49 Sustainable forest management demands accurate, efficient and up-to-date  
50 information in order to describe forest structure and quantify forest productivity (Burkhart  
51 and Tomé, 2012). As forest ecosystems are dynamic, it is essential to simulate forest stand  
52 attributes projected temporally and spatially to support forest management (Tompalski et  
53 al., 2016). Site productivity and forest growth are critical inputs for projecting wood  
54 volume and biomass accumulation over time (Skovsgaard and Vanclay, 2008; Diéguez-  
55 Aranda et al., 2012). Site productivity, which is commonly determined using site index  
56 (SI) models, is also the primary option for many forest management decisions (Álvarez-  
57 González et al., 2005; Diéguez-Aranda et al., 2006a, 2009, 2012).

58 Although SI may not fully represent true site productivity, it is the method most widely  
59 used to estimate site productivity for even-aged forests (Álvarez-González et al., 2005;  
60 Diéguez-Aranda et al., 2006b). Theoretical growth curves have been developed for  
61 estimating SI and predicting dominant height growth ( $\Delta H$ ) of the main productive species  
62 in Galicia (for further details, see Diéguez-Aranda et al. (2012). These curves are based on  
63 biological relationships and thus give specific meaning to the parameters (e.g. maximum  
64 size, growth rate). The curves predict different variables as a function of age, identifying  
65 exponential, linear and asymptotic growth stages throughout the lifespan of the tree (Arias-  
66 Rodil et al., 2015). However, stand age may not be available for inclusion in growth  
67 models. Measurement of this variable can be difficult, as it requires extraction of whole  
68 increment cores in tree species with well-defined tree rings (Tomé et al., 2006). The  
69 aforementioned study proposed a method for formulating growth functions as age-  
70 independent difference equations. Following this approach, Arias-Rodil et al. (2015)  
71 concluded that an age-independent equation depending on climatic factors performed best  
72 for predicting  $\Delta H$ , outperforming the traditional age-dependent site quality curves. Recent  
73 studies (Scolforo et al., 2019, 2020; González-Rodríguez and Diéguez-Aranda, 2020)  
74 indicate the importance of improving the existing growth models by incorporating  
75 environmental variables developing site quality maps for different species. Consequently,  
76 further research is required to deliver a compatible set of growth equations for prediction  
77 and projection. Such equations should include parameters refined by environmental  
78 variables to make the empirical models spatially applicable under climate change scenarios.

79 A number of researchers have pointed out various desirable attributes of dominant  
80 height ( $H$ ) growth models (e.g. Bailey and Clutter, 1974; Clutter et al., 1983; Cieszewski  
81 and Bailey, 2000), including the following: (i) consistency, i.e. no change for zero elapsed  
82 time; (ii) path-invariance, whereby the results of projecting the dominant height first from  $t_0$   
83 to  $t_1$ , and secondly from  $t_1$  to  $t_2$ , must be the same as those of the one-step projection from  $t_0$   
84 to  $t_2$ ; (iii) causality, in which a change in dominant height can only be influenced by inputs  
85 within the relevant time interval; (iv) polymorphism, i.e. curves for different SI values  
86 should not be proportional; (v) a sigmoid growth pattern with an inflexion point; (vi),  
87 horizontal asymptote at old ages, (vii) logical behaviour (height should be zero at age zero  
88 and equal to SI at the reference age and the curve should never decrease); and (viii)  
89 theoretical basis or interpretation of model parameters derived by analytically tractable  
90 algebraic operations and base-age invariance. Fulfilment of the aforementioned  
91 requirements depends on both the construction method and the specific mathematical

92 function. The algebraic difference approach (ADA) developed by (Bailey and Clutter,  
93 1974), and its generalization, known as GADA (Cieszewski and Bailey, 2000), are useful  
94 for this purpose. Numerous studies have used GADA to model the dominant height growth  
95 or site index in pine forests in NW Spain (Nunes et al., 2011; Álvarez-González et al.,  
96 2005), South of Europe (Bravo-Oviedo et al., 2007, 2008; Kahriman et al., 2018), Central  
97 Europe (Socha et al., 2020) and North and South America (Lauer and Kush, 2010; Vargas-  
98 Larreta et al., 2013; Scolforo et al., 2016). Dominant height growth ( $\Delta H$ ) equations have  
99 the general form (omitting the vector of model parameters) of  $Y = f(t, t_0, Y_0)$ , where  $Y$  is the  
100 dominant height at age  $t$ , and  $Y_0$  is the reference variable defined as the value of the  
101 dominant height at age  $t_0$ . The ADA essentially involves replacing a base-model site-  
102 specific parameter with its initial-condition solution. The GADA allows expansion of the  
103 base equations according to various theories about growth characteristics (e.g. asymptote,  
104 growth rate), thereby allowing more than one parameter to be site-specific and allowing  
105 derivation of more flexible dynamic equations (see Cieszewski, 2002).

106 In Spain, the only forest growth information currently available at national and regional  
107 scales is that provided by the Spanish National Forest Inventory (SNFI), which began in  
108 1965 (SNFI1). Since the second round of measurements (SNFI2, 1986), the SNFI has been  
109 based on a systematic sampling design with permanent plots and previous forest  
110 stratification (Spanish Forest Map, SFM) determined by photo-interpretation. It is  
111 important to note that the SNFI showcase (Fernández-Landa et al., 2018; Pascual et al.,  
112 2020) requires upgrading of the ground positioning protocols to improve data co-  
113 registration. Ongoing efforts in this regard (SNFI4 and SNFN, i.e. SNFIs of productive  
114 stands in northern Spain) are reducing the gap between nominal and true positions for SNFI  
115 samples to allow joint use of the SNFI and remote sensing data for spatial estimation of  
116 forest attributes (Pascual et al., 2020).

117 Airborne laser scanning (ALS) has been a primary source of 3D data on forest vertical  
118 structure since the 1990s (Naesset, 1997; Maltamo et al., 2014). An abundance of research  
119 has demonstrated the utility of ALS for predicting forest biophysical variables to support  
120 forest inventories at individual tree- and stand- level (Hyypä et al., 2008), providing  
121 accurate estimates of tree and stand height (Gatziolis et al., 2010; Guerra-Hernández et al.,  
122 2016; Maltamo et al., 2014; Roussel et al., 2017). Multitemporal ALS data have also been  
123 used to measure forest height growth, with no statistically significant differences between  
124 the field- and ALS-derived mean height increment measurements (Yu et al., 2004, 2006;  
125 Hopkinson et al., 2008). A number of recent studies have shown that ALS or image-based

126 point clouds (obtained by stereo aerial photography) may be valid for estimating SI (Véga  
127 and St-Onge, 2008; Tompalski et al., 2015; Socha et al., 2017; Solberg et al., 2019;  
128 Noordermeer et al., 2018; 2020; Socha et al., 2020). For instance, Noordermeer et al. (2018,  
129 2020) derived SI based on bi-temporal ALS surveys and existing stand-level models or  
130 Solberg et al. (2019) developed an age-independent approach to estimate SI for spruce-  
131 dominated stands using repeated single-tree ALS data.

132 Most studies on ALS and SI have relied on existing height growth models, developed  
133 with data collected from permanent sample plots or stem analysis. Calibration of SI models  
134 can be replaced by site-specific height growth equations based on growth detected between  
135  $t_1$  and  $t_2$  (the time lapse between ALS surveys), as Socha et al. (2017) reported for Norway  
136 spruce forests. The stages of development of commercial tree species are well known, and  
137 models based on height dynamics using repeated ALS may therefore be recognized as new,  
138 fully valuable wall-to-wall data sources for tree height growth and SI modelling (Socha et  
139 al., 2020). In addition, multi-temporal ALS data may also be used for rapid, accurate and  
140 cost-effective tree growth assessment, providing up-to-date information to support  
141 decision-making in forest management (Hopkinson et al., 2008; Tompalski et al., 2016,  
142 2018, 2019).

143 The main objective of this study was to predict ( $\Delta H$ ) using an age-independent method  
144 based on the GADA approach and data from two ALS acquisitions. The work involved the  
145 following: (i) comparison of two GADA models based on two well-known base models, to  
146 predict and project  $H$  and SI in *P. pinaster* stands in Galicia from data acquired in two ALS  
147 flights and (ii) establishment of a Multivariate Adaptive Regression Splines (MARS)  
148 relationship between the site-specific parameter of the GADA model and environmental  
149 variables to enable use of the model when data from two ALS flights is not available.

150

## 151 **Materials and Methods**

### 152 *Study area*

153 The study area is located in Galicia, in north-west Spain (Fig. 1). This region is  
154 characterized by rugged orography and an oceanic climate with mild temperatures, low  
155 thermic oscillation between winter and summer and frequent rainfall. The region, which has  
156 1.4 million hectares of tree-covered land and more than 68 million cubic metres of standing  
157 timber, provides 51% and 27% of respectively the conifer and total annual harvest volume  
158 in Spain (MAGRAMA, 2010). *Pinus pinaster* Ait. is one of the most widely distributed  
159 forest tree species in this region, covering an area of more than 276 000 ha (MAPA, 2018).

160

161 *Field data*

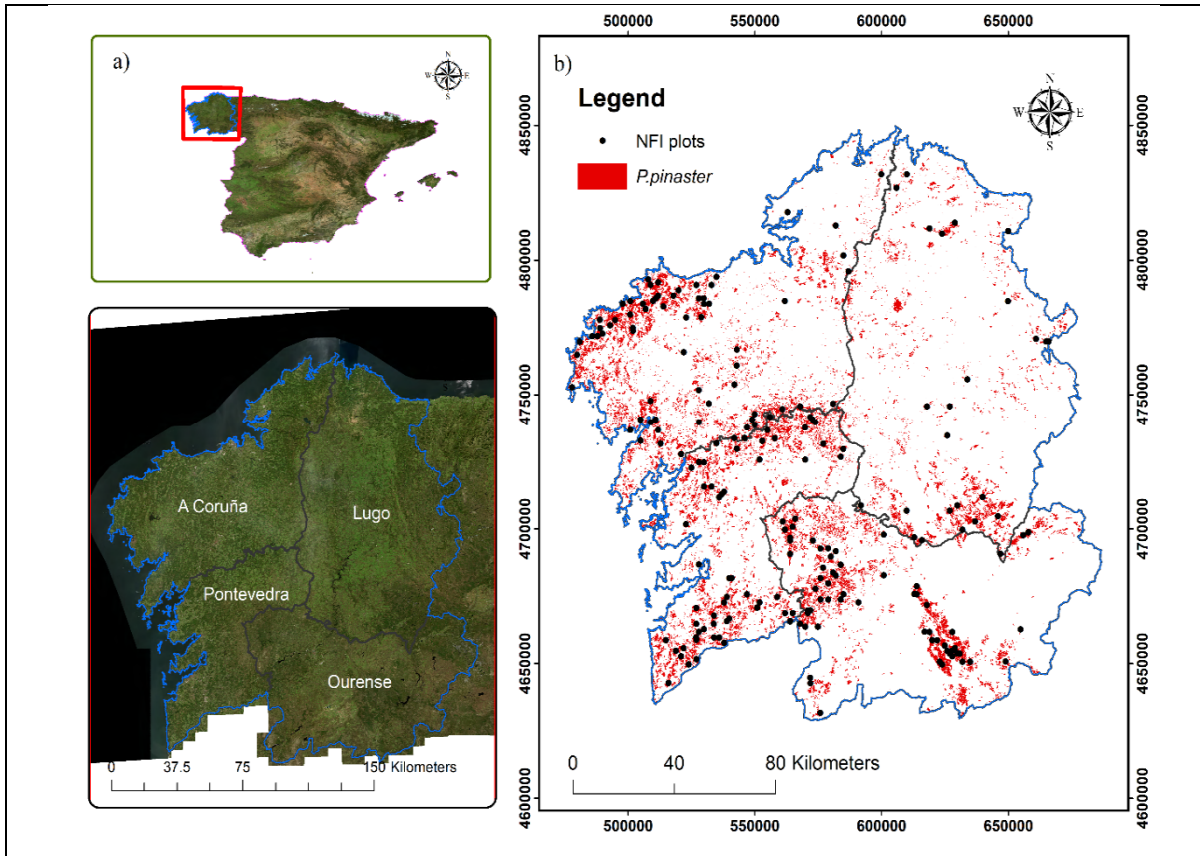
162 The field data used for this study were obtained from the SNFI-4 and SNFI-N. The  
163 field measurements of the SNFI-4 in Galicia data were carried out between March and  
164 September 2009 (MAGRAMA, 2012), whereas field measurement for the SNFI-N were  
165 carried out between May and October 2018 (MAGRAMA, 2018). The SNFI project  
166 maintains a network of sample plots throughout the whole country, which are aimed at  
167 providing continuously up-dated information on the status of nationwide forest resources,  
168 including timber volumes and species composition. Sample plots are established at the  
169 intersections of a 1 x 1 km UTM grid and consist of four circular concentric subplots of  
170 radius 5, 10, 15 and 25 m. Trees are selected depending on the distance to the plot centre  
171 and the diameter at breast height (1.3 m, *dbh*): *i*) trees with  $7.5 \leq dbh < 12.5$  cm were  
172 measured within a subplot radius of 5 m, *ii*) trees with  $12.5 \leq dbh < 22.5$  cm were measured  
173 within a subplot radius of 10 m, *iii*) trees with  $22.5 \leq dbh < 42.5$  cm were measured within  
174 a subplot radius of 15 m and *iv*) trees with a  $dbh \geq 42.5$  cm were measured within a subplot  
175 radius of 25 m (Álvarez-González et al., 2014; Alberdi et al., 2017). For all trees in the  
176 plots, the total height (*h*) was measured with a hypsometer, to the nearest 0.1 m, and the  
177 *dbh* was measured in two perpendicular directions with a caliper, to the nearest 0.1 cm. The  
178 following stand variables were calculated from the tree variable measurements by using  
179 tree expansion factors: number of stems per hectare (*N*), quadratic mean diameter (*dg*),  
180 stand basal area (*G*), mean stand height (*h*) and dominant height (*H*, defined as the mean  
181 height of the 100 thickest trees per ha). The expansion factor can be defined as the  
182 relationship between the reference area (1 ha) and the subplots area adjusting the values of  
183 the number of sampled trees to a per hectare value (Álvarez-González et al., 2014)

184 As stand age is not measured in the SNFI, the site index (SI) of the sample plots cannot  
185 be directly estimated from the SI curves developed for the species in the study area.  
186 However, an indirect method was used to estimate stand age and SI for SNFI-4 and SNFI-N  
187 data. This method requires two measurements of dominant height ( $H_{SNFI-4}$  and  $H_{SNFI-N}$ ) at  
188 ages  $t_1$  and  $t_1 + \Delta t$  for each sample plot, where  $\Delta t$  is the time interval between SNFI-4 and  
189 SNFI-N. Using these values, the algebraic difference equation of the SI curves of the  
190 species in Galicia (Diéguez-Aranda et al., 2009) can be solved numerically to estimate  $t_1$ .  
191 Once  $t_1$  is obtained, the same equation can be used to estimate *SI* (m) at a reference age of  
192 20 years, from the values of  $H_{SNFI-4}$  and  $t_1$ .

193 During field measurement of SNFI-N, the co-registration uncertainty was addressed by  
194 using commercial grade global navigation satellite systems (GNSS) to upgrade positioning  
195 information on new samples established in forest areas under expansion and on the set of  
196 previously existing SNFI-4 sample plots. A handheld data collection system (TRIMBLE  
197 Juno 5B handheld, Trimble Inc. USA) was used to determine the coordinates (error range  
198 1-2 m of positioning error after post-processing).

199 In the first step, all sample plots in which *P. pinaster* predominated (i.e. it represented  
200 more than 90% of the trees per ha and more than 90% of total stand basal area in the plot)  
201 in both SNFI-4 and SNFI-N were selected. In the second step, exhaustive examination of  
202 the data was carried out to reject sample plots with atypical dominant height growth, i.e.  
203 values outside the range observed in the *SI* curves of the species in the study area (Dieguez-  
204 Aranda et al., 2009); sample plots with lower dominant height and sample plots with dead  
205 trees accounted for more than 10% of the basal area. Finally, a total of 156 sample plots of  
206 *P. pinaster* were selected for the purposes of the study.

207 Figure 1 shows the spatial distribution of pure and even-aged *P. pinaster* stands and the  
208 selected SNFI sample plots in Galicia. The main stand variables of the SNFI-4 and SNFI-N  
209 sample plots used in this study are summarized in Table 1.



210 Figure 1. (a) Location of the study area. (b) Spatial distribution of the Spanish SNFI-4  
 211 and SNFI-5 plots where *P. pinaster* was dominant (black dots) and the spatial  
 212 distribution of pure *P. pinaster* (red) stands from latest version of Spanish Forest Map  
 213 (MAPA, 2018).

214

215 Table 1. Summarized statistics of the stand variables observed within 156 field plots.

Inventory	Variable	Mean	Std. dev.	Maximum	Minimum
SNFI-4	<i>N</i>	747.85	476.54	2300.46	114.31
	<i>G</i>	28.18	12.19	64.16	8.02
	<i>H</i>	17.47	4.78	28.85	6.33
	<i>V</i>	199.57	112.36	550.15	27.01
SNFI-N	<i>N</i>	591.15	374.64	2038.74	110.15
	<i>G</i>	33.62	12.44	80.15	13.52
	<i>H</i>	21.49	4.61	32.62	10.74
	<i>V</i>	273.25	124.32	727.60	37.99
SNFI-4 SNFI-N	<i>SI</i>	15.67	5.33	29.81	5.60

216 where *N* is the stand density (trees ha<sup>-1</sup>), *G* is the stand basal area (m<sup>2</sup> ha<sup>-1</sup>), *H* is the  
 217 dominant height (m), *V* is the stand volume (m<sup>3</sup> ha<sup>-1</sup>), and *SI* is the site index (m), i.e. the  
 218 value of dominant height at a reference age of 20 years.

219

#### 220 *ALS acquisition and processing*

221 Two sets of ALS point clouds were processed in this study: the ALS data for the first  
 222 coverage were collected in the periods September to October 2009 (Galicia East) and  
 223 February to April 2011 (Galicia West), while the ALS survey in the second coverage took  
 224 place in the periods July-September 2015 (Galicia West), August-December 2016 and  
 225 February 2017 (Galicia East). Both data sets correspond to the first and second round of  
 226 countrywide ALS measurements, which are publicly available in Spain through the  
 227 National Plan for Aerial Orography (hereafter referred to as PNOA). Square ALS blocks of  
 228 2 km side, covering the whole region of Galicia were obtained from the CNIG (Centro  
 229 Nacional de Información Geográfica) computer server  
 230 (<http://centrodedescargas.cnig.es/CentroDescargas/index.jsp>). The scanning sensors  
 231 involved in collecting the ALS data in the study area were a RIEGL LMS-Q680 for the  
 232 whole first coverage, and a LEICA ALS50 (West Galicia) and a LEICA ALS80 (East

233 Galicia) for the second coverage. The nominal laser pulse density varied between 0.5 first  
234 returns  $\text{m}^{-2}$  in the first coverage and 0.5-1 first returns  $\text{m}^{-2}$  in the second coverage. The  
235 vertical accuracy of scanning survey varied between 0.2 and 0.4 m for the first coverage  
236 and 0.15 and 0.2 m for the second coverage. The ALS data sets were processed using  
237 Lastools software (Isenburg, 2020). A detailed description of the software parametrization  
238 and all the processing workflow of ALS point cloud is given by Pascual et al., (2020).  
239 Briefly, all ALS echoes classified as ground were used to create a digital elevation model  
240 (DEM) considering a spatial resolution of 2 m. The 99<sup>th</sup> percentile of height distribution  
241 was extracted from the normalized ALS point cloud, using a buffer of 25 m of radius from  
242 the SNFI-N coordinates of the plot centres (GNSS-based measurements). The above-  
243 ground height of ALS echoes was used to distinguish tree canopies (echoes above 2 m) and  
244 the shrub layer (echoes below 2 meters) when computing the ALS height statistics  
245 (*lascanopy* parameters: `height_cutoff = 2`, `cover_cutoff = 2`).

246

#### 247 *Physiographical, edaphic and climate data*

248 The 2-m resolution DEM was used as the source of physiographical data. At plot level,  
249 the ALS ground points were converted to a triangular irregular networks (TIN) surface  
250 raster, and 2-m resolution raster using memory-efficient-streaming technology under three  
251 parallel processes (Isenburg 2020). The following topographic variables were calculated in  
252 R software (R Core Team, 2020), using the *terrain* function in the R *raster* package  
253 (Hijmans et al., 2015): elevation, slope, aspect, topographical position index (*TPI*), terrain  
254 roughness index (*TRI*) and roughness. Moreover, two trigonometric transformations of  
255 azimuth aspect were calculated to characterize the exposure to north-south (*cos $\omega$* ) and east-  
256 west effects (*sin $\omega$* ).

257 The soil variables were obtained from a 500-m resolution raster developed by (Ballabio  
258 et al., 2016) for the European territory. These maps provide a set of variables related to soil  
259 physical attributes, estimated through interpolation from the soil samples of the LUCAS  
260 2009 TOPSOIL database of the European Soil Data Center (ESDAC) (Panagos et al.,  
261 2012). Six soil variables were selected from these raster maps as potential predictors of *SI*:  
262 the percentages of sand (*%Sand*), silt (*%Silt*), clay (*%Clay*) and coarse material (*%Coarse*),  
263 the soil bulk density (*SBD*,  $\text{Mg m}^{-3}$ ), and the available water capacity (*AWC*), defined as the  
264 difference between the field capacity and the water content at permanent wilting point.

265 Finally, the Digital Climate Atlas for the Iberian Peninsula (Ninyerola et al., 2000,  
266 2007; Pons and Ninyerola, 2008), with 200-m spatial- and monthly temporal- resolution,

267 was used as a source of climate data. The climate variables included in the dataset are  
268 average monthly temperature, average maximum and minimum monthly temperature and  
269 monthly total precipitation. A set of annual and seasonal climate variables were derived  
270 from these original data (<http://opengis.uab.es/wms/iberia/>). In addition (Table 2), the  
271 potential evapotranspiration ( $Pe$ ) was calculated using Thornthwaite's (1933) equation, and  
272 the compensated thermicity index ( $Cti$ ) was obtained using the equation developed by  
273 Rivas-Martínez (2008).

274 The descriptive statistics of the physiographical, edaphic and climate variables were  
275 extracted from the raster maps for the specific locations of the 156 previously selected  
276 sample plots, thus considering its constancy throughout development of each stand. The  
277 physiographical, edaphic and climate variables are summarized in Table 2.

278 Table 2. Summary statistics of the physiographical, edaphic and climate variables extracted from the raster datasets corresponding to the 156 sample  
 279 plot locations.

Variable	Variable	Description	Mean	Std. dev.	Maximum	Minimum
<i>Physiographical</i>	<i>elevation</i> (m)	Orthometric elevation	359.26	228.93	912.38	33.88
	<i>slope</i> (%)		12.62	7.35	35.13	0.65
	<i>cos<math>\omega</math></i>	North-south effect of azimuth ( $\omega$ )	-0.03	0.70	1.00	-1.00
	<i>sin<math>\omega</math></i>	East-west effect of azimuth ( $\omega$ )	-0.05	0.71	1.00	-1.00
	<i>TPI</i> (m)	Topographic position index	0.01	0.26	0.85	-1.05
	<i>TRI</i> (m)	Terrain roughness index	1.78	1.10	5.62	0.09
	<i>Roughness</i> (m)		5.77	3.66	17.99	0.55
<i>Edaphic</i>	<i>%Sand</i>	Sand textural percentage	55.99	9.25	75.03	30.61
	<i>%Silt</i>	Silt textural percentage	32.28	7.28	52.83	17.56
	<i>%Clay</i>	Clay textural percentage	11.73	2.76	24.63	7.00
	<i>%Coarse</i>	Percentage coarse matter	21.88	6.63	40.17	8.86
	<i>SBD</i> (Mg m <sup>-3</sup> )	Soil bulk density	1.12	0.03	1.16	1.01
	<i>AWC</i>	Available water capacity (soil volume fraction)	0.093	0.006	0.11	0.08
<i>Climate</i>	<i>T</i> (°C)	Mean annual temperature	12.74	1.13	14.99	10.10
	<i>P</i> (mm)	Mean annual precipitation	1280.23	126.53	1529.73	779.58
	<i>T<sub>max</sub></i> (°C)	Mean annual maximum temperature	18.04	1.19	20.60	14.90
	<i>T<sub>min</sub></i> (°C)	Mean annual minimum temperature	7.451	1.37	9.90	4.70
	<i>T<sub>wm</sub></i> (°C)	Mean temperature of warmest month	19.42	1.20	21.7	15.33

$T_{cm}$ (°C)	Mean temperature in the coldest month	7.12	1.43	9.42	3.68
$M$ (°C)	Mean maximum temperature in the warmest month	25.95	1.88	28.95	19.59
$m$ (°C)	Mean minimum temperature in the coldest month	2.96	1.48	5.85	-0.61
$T_s$ (°C)	Mean summer temperature (June, July and August)	18.51	1.65	21.70	13.30
$P_s$ (mm)	Mean summer precipitation (June, July and August)	129.34	24.06	187.54	77.46
$Pe$ (mm)	Potential evapotranspiration (Thornthwaite, 1933)	963.24	123.16	1251.33	706.02
$Ci$ (°C)	Continental interval ( $T_{max} - T_{min}$ )	22.99	2.71	27.41	17.2
$Cti$ (°C)	Compensated thermicity index (Rivas-Martinez, 2008)	371.64	48.66	464.68	258.63

280 where  $TPI = y - \sum_{i=1}^8 \frac{y_i}{8}$ ;  $TRI = \sum_{i=1}^8 \frac{|y_i - y|}{8}$ , with  $y$  is the altitude of a cell and  $y_i$  is the altitude of the  $i$ -th adjacent cell;  $Pe = \sum_{i=1}^{12} 16 \left( \frac{10T_i}{I} \right)^a$ , with  
281  $I = \sum_{i=1}^{12} \left( \frac{T_i}{5} \right)^{1.514}$ ,  $a = 0.49 + 0.079 \cdot I - 7.71 \cdot 10^{-5} \cdot I^2 + 6.75 \cdot 10^{-7} \cdot I^3$  and  $T_i$  is the mean temperature of the  $i$ -th month;  $Cti = 10(T +$   
282  $T_{max} + T_{min}) - 15 \cdot (Ci - 20)$ .

283 *Relationship between dominant height and ALS data*

284 In the present study, the *SI* curves were derived using the generalized algebraic  
285 difference approach (GADA). The model was developed from the estimates of *H* obtained  
286 from the LiDAR metrics of each sample plot and field inventory (SNFI-4 and SNFI-N). For  
287 this purpose, a relationship between the *H* field measurements and the LiDAR metrics was  
288 previously fitted.

289 Once the *H* and age values of each sample plot in each field inventory were estimated,  
290 the *H* of the sample plots at the time of LIDAR data acquisition was estimated using the  
291 site quality curves for this species (Diéguez-Aranda et al., 2009) and the time interval  
292 between field measurement dates and flight dates.

293

294 *Dominant height growth model approach*

295 Two different base equations were evaluated for modelling the  $\Delta H$  pattern using the  
296 GADA approach: Von Bertalanffy, 1957 and Korf, 1939. These functions have been widely  
297 used in modelling *H* (e.g., Monserud, 1984; Cieszewski, 2002; Álvarez-González et al.,  
298 2005; Diéguez-Aranda et al., 2006a; Castedo-Dorado et al., 2007; Socha et al., 2020).

299 As general notational convention,  $a_1, a_2, \dots, a_n$  were used to denote the parameters in  
300 base models, while  $b_1, b_2, \dots, b_m$  were used for global parameters in subsequent GADA  
301 formulations.

302 The first GADA formulation was tested using the base model proposed by Bertalanffy  
303 (1957), the integral form of which can be represented, for the sake of simplicity, as

$$H = a_1(1 - \exp(-a_2t))^{a_3} \quad [1]$$

304 where  $t$  is the stand age,  $a_1$  is an asymptote or limiting value,  $a_2$  is often called a ‘rate  
305 parameter’, and  $a_3$  is often referred to as ‘an initial pattern parameter’. To derive models  
306 with both polymorphism and variable asymptotes from Equation (1), more than one  
307 parameter must be site specific. Thus, both the asymptote  $a_1$  and the shape parameter  $a_3$  are  
308 assumed to be dependent on  $X$ , a site-specific (local) parameter, using the relationships  
309 proposed by Cieszewski (2004):

$$a_1 = \exp(X) \text{ and } a_3 = b_2/X, \text{ while } a_2 = b_1$$

310 which defines the following relationship:

$$H = \exp(X)(1 - \exp(-b_1t))^{b_2/X} \quad [2]$$

311 Considering two measurements of the same sample plot ( $t_1, H_1$ ) and ( $t_2, H_2$ ), both point  
312 correspond to the same *SI* curve with the same values of local ( $X$ ) and global parameters ( $b_1$   
313 and  $b_2$ ):

$$\begin{aligned}
314 \quad & \left. \begin{aligned} H_1 &= \exp(X)(1 - \exp(-b_1 t_1))^{b_2/X} \\ H_2 &= \exp(X)(1 - \exp(-b_1 t_2))^{b_2/X} \end{aligned} \right\}
\end{aligned}$$

315 Solving  $t_1$  in the first expression and substituting this value in the second one, while  
316 considering that  $t_2 = t_1 + \Delta t$ , yields:

$$\left. \begin{aligned} H_1 &= \exp(X)(1 - \exp(-b_1 t_1))^{b_2/X} \\ H_2 &= \exp(X)(1 - \exp(-b_1 t_2))^{b_2/X} \end{aligned} \right\} \Rightarrow t_1 = -\log \left( 1 - \left( \frac{H_1}{\exp(X)} \right)^{X/b_2} \right) / b_1$$

317 We can then obtain the first GADA formulation [E1] with the general expression  $H_2 =$   
318  $f(H_1, \Delta t)$  and with two global parameters ( $b_1$  and  $b_2$ ) and a site-specific (local) parameter  
319 ( $X$ ):

$$H_2 = \exp(X) \left( 1 - \exp \left( -b_1 \left[ \Delta t - \log \left( 1 - \left( \frac{H_1}{\exp(X)} \right)^{X/b_2} \right) / b_1 \right] \right) \right)^{b_2/X} \quad [E1]$$

320 The second GADA formulation is based on the base model proposed by Korf (1939):

$$H = a_1 \exp(-a_2 t^{-a_3}) \quad [3]$$

321 where  $a_1$  is the asymptote parameter, and the other two parameters ( $a_2$  and  $a_3$ ) are related to  
322 the inflexion point and the growth rate. Cieszewski, 2002 replaced  $a_1$  with the exponential  
323 of the unobserved site-specific variable  $X$ , and  $a_2$  by  $b_1/X$ , while  $a_3 = b_2$

$$H = \exp(X) \exp(-(b_1/X)t^{-b_2}) \quad [4]$$

324 Again, considering two measurements of the same sample plot and solving for  $t_1$  and  
325 substituting in the second expression, we obtained the second GADA formulation:

$$\begin{aligned}
326 \quad & \left. \begin{aligned} H_1 &= \exp(X) \exp(-(b_1/X)t_1^{-b_2}) \\ H_2 &= \exp(X) \exp(-(b_1/X)t_2^{-b_2}) \end{aligned} \right\} \Rightarrow t_1 = \left[ -\log \left( \frac{H_1}{\exp(X)} \right) \left( \frac{X}{b_1} \right) \right]^{-1/b_2} \\
& H_2 = \exp(X) \exp \left( -(b_1/X) \left( \Delta t + \left[ -\log \left( \frac{H_1}{\exp(X)} \right) \left( \frac{X}{b_1} \right) \right]^{-1/b_2} \right)^{-b_2} \right) \quad [E2]
\end{aligned}$$

327

### 328 *Dominant height growth model fitting*

329 The GADA models [E1] and [E2] were fitted, by using the SAS/ETS<sup>®</sup> MODEL  
330 procedure (SAS Institute Inc., 2004), to the complete database of sample plots to obtain a  
331 single guide-curve with the same global parameters ( $b_1$  and  $b_2$ ) and a unique estimate of  $X$ ,  
332 common to all sample plots.

333 This guide curve is used as a reference to obtain the specific  $\Delta H$  curve of a new sample  
334 plot ( $i$ ); for this purpose, the first step consists of estimating the local parameter  $X_i$  using  
335 [E1] or [E2] with the global parameters of the respective guide-curve and the measurements

336 of  $H_1$ ,  $H_2$  and  $\Delta t$  for that specific sample plot, i.e.  $X_i = f(b_1, b_2, H_{1i}, H_{2i}, \Delta t_i)$ . Once the  
 337 value of parameter  $X_i$  is calculated, the specific curve for the sample plot is obtained using  
 338 equation [2] or equation [4], with the global parameters of [E1] or [E2], respectively and  
 339 the local, previously estimated parameter.

340 In this study, the value of  $X$  of each of the sample plots was calculated with the  
 341 *optimize* function of R (R Core Team, 2020), which uses the method of Brent (1973)) and is  
 342 appropriate for one-dimensional optimization.

343 Comparison of the estimates for the two different GADA models was based on  
 344 numerical and graphical analysis of the residuals ( $e_{ij}$ ). Two statistical criteria were then  
 345 examined: the root mean square error (*RMSE*), which analyses the accuracy of the  
 346 estimates, and the model efficiency (*MEF*), which shows the proportion of the total  
 347 variance explained by the model, adjusted for the number of model parameters and the  
 348 number of observations. *MEF* compares predictions directly with observed data using a  
 349 statistic analogous to adjusted  $R^2$  (Vanclay and Skovsgaard, 1997). This statistic provides a  
 350 simple index of performance on a relative scale, where 1 indicates a ‘perfect’ fit and 0  
 351 reveals that the model is no better than a simple average. The expressions of these statistics  
 352 are as follows:

$$RMSE = \sqrt{\frac{\sum_{i=1}^n (y_i - \hat{y}_i)^2}{n - p}}$$

$$MEF = 1 - \frac{(n - 1) \sum_{i=1}^n (y_i - \hat{y}_i)^2}{(n - p) \sum_{i=1}^n (y_i - \bar{y})^2}$$

353 where  $y_i$ ,  $\hat{y}_i$  and  $\bar{y}$  are the observed, predicted and average values of the dominant height,  
 354 respectively;  $n$  is the total number of observations used to fit the function; and  $p$  is the  
 355 number of model parameters.

356 Other important steps in evaluating the functions fitted include graphical analysis of  
 357 the residuals and examination of the appearance of the fitted curves overlaid on the  
 358 trajectories of the dependent variables for each plot. Visual or graphical inspection is an  
 359 essential point in selecting the most appropriate model, because curve profiles may differ  
 360 drastically, even when the fitting statistics and residuals are similar (Huang et al., 2003).

361

362 *Estimation of the site-specific parameter (X) from physiographical, edaphic and climate*  
 363 *data*

364

365 The GADA models [E1] and [E2] can only be used to obtain the dominant height  
 366 growth curve of a new sample plot or stand when data from two LiDAR flights are  
 367 available to estimate the site-specific parameter  $X$  ( $X_i = f(b_1, b_2, H_{1i}, H_{2i}, \Delta t_i)$ ). However,  
 368 from a practical point of view, obtaining this curve from a single flight would be very  
 369 useful. As parameter  $X$  is a specific indicator of productivity because it affects both the  
 370 asymptote of the site quality curve and its growth pattern, an explicit relationship between  
 371 this parameter and the site environmental conditions, defined by the physiography, soil and  
 372 climate variables, was evaluated.

373 The modelling approaches used should enable development of parsimonious models,  
 374 preventing overfitting and resulting in relationships where ecological coherence can be  
 375 easily assessed. In this study, the Multivariate Adaptive Regression Splines (MARS)  
 376 approach was used to model this relationship. This approach was selected on the basis of  
 377 the results obtained by González-Rodríguez and Diéguez-Aranda (2020), who evaluated  
 378 seven different parametric and non-parametric methods for relating the site index of *Pinus*  
 379 *radiata* D. Don. stands in the study region (Galicia) to environmental variables, and  
 380 proposed this method as the most suitable approach.

381 MARS solution is a nonparametric technique proposed by Friedman, 1991, based on  
 382 fitting piecewise linear regressions by intervals of the independent variable space. The  
 383 general form of a MARS model is as follows:

$$y = f_M(x) + \varepsilon \quad [5]$$

384 where  $\varepsilon$  is the error, and  $f_M(x)$  is the unknown regression function, derived as follows:

$$f_M(x) = \beta_0 + \sum_{i=1}^M \beta_i B_i(x) \quad [6]$$

385 where  $\beta_0$  is the intercept of the model,  $B_i(x)$  are piecewise linear basis functions,  $\beta_i$  is the  
 386 coefficient of the  $i^{\text{th}}$  base, and  $M$  is the number of base functions. Each piecewise linear  
 387 base function takes on the following two forms: 1) a hinge function with the form  $\max(0, x-$   
 388  $k)$  with  $x-k > 0$  or  $\max(0, k-x)$  with  $k-x > 0$  where  $k$  is a constant value called knot; or 2) a  
 389 product of two or more hinge functions that can therefore model interactions between two  
 390 or more independent variables ( $x$ ) defining terms of first, second, third degree, etc.

391 The optimal MARS model is determined in a two-stage process. First, MARS  
 392 constructs a very large number of base functions that fit the data with different degrees of  
 393 accuracy. Second, the generalized cross-validation (GCV) criterion is used and the set of  
 394 base functions with the lowest value of GCV is selected:

$$\text{GCV}(M) = \frac{\sum_{i=1}^n (y_i - f_M(x_i))^2}{n \left(1 - \frac{p_M}{n}\right)^2} \quad [7]$$

395 where  $y_i$  represents the observed values of dominant height,  $n$  is the number of sample plots  
 396 and  $p_M$  is the number of model parameters. The MARS model was fitted with the earth  
 397 package (Milborrow et al., 2019) implemented in R software (R Core Team, 2020).

398 The use of these large-scale models outside the range of values of the sample used to  
 399 develop them could result in abnormal estimates of the local parameter  $X$ , which would  
 400 produce unusual site index and dominant height growth values. To avoid this drawback,  
 401 some restrictions have been established for the maximum and minimum estimates of this  
 402 parameter  $X$ . In a comprehensive assessment of the site quality of the species in the study  
 403 area, the minimum and maximum  $SI$  values derived from stem analysis of 204 dominant  
 404 trees were respectively 6.7 and 21.9 m (Álvarez-González et al., 2005). Therefore the  
 405 extreme estimates of the  $X$  parameter will be those leading to these maximum and  
 406 minimum of the site index.

407

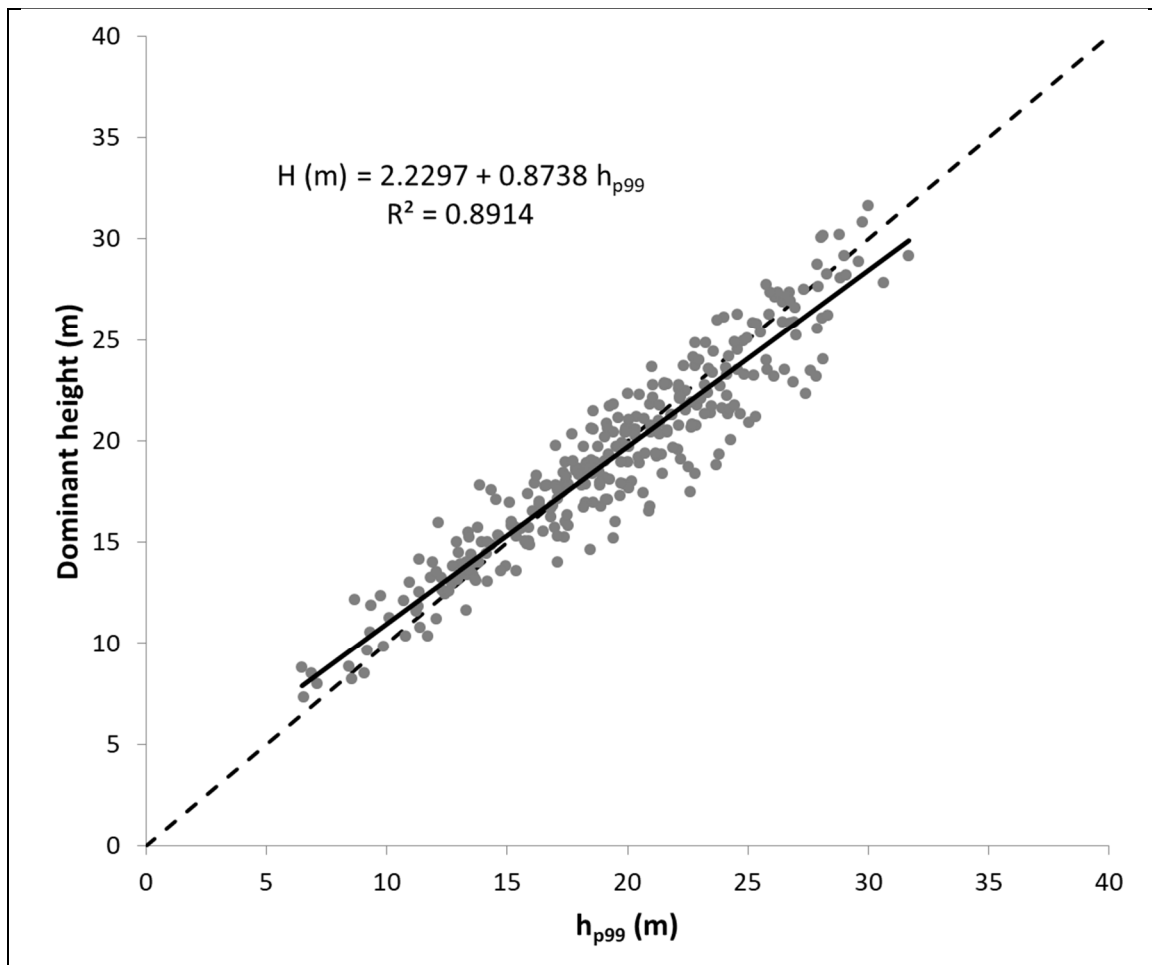
## 408 **Results**

### 409 *Relationship between dominant height and ALS data*

410

411 The Pearson's correlation coefficients were computed for  $H$  field measurements and  
 412 the ALS percentiles. The highest value was obtained for the 99<sup>th</sup> percentile, and a linear  
 413 relationship for estimating  $H$  from  $h_{p99}$  was thus fitted using the previously estimated values  
 414 of  $H$  for the 156 sample plots at the time of both LiDAR acquisitions. The model explained  
 415 more than 89% of the observed variability (Figure 2). Finally, the fitted relationship was  
 416 used to estimate the dominant heights of each sample plot for the two LiDAR acquisition  
 417 dates.

418



419 Figure 2. Linear relationship and coefficient of determination ( $R^2$ ) between estimated  
 420 dominant height at the time of both LiDAR acquisitions ( $H$ , m) and the 99th percentile  
 421 heights from the distribution of ALS echoes.

422

423 *Dominant height growth model*

424 The parameter estimates for both GADA equations [E1] and [E2], the approximate  
 425 standard errors and the corresponding goodness-of-fit statistics are shown in Table 3. Both  
 426 GADA equations produced very similar values for the goodness-of-fit statistics, explaining  
 427 more than 91% of the total observed variability in height increment. In addition, the  
 428 distribution of residuals against the estimated values is also very similar in both models  
 429 (Figure 3). For a more detailed comparison between the models, the mean annual increment  
 430 (MAI) in dominant height in each plot has been estimated for each model (E[1] and E[2])  
 431 and compared with the field-observed values for different age classes and site indexes  
 432 (Figure 4).

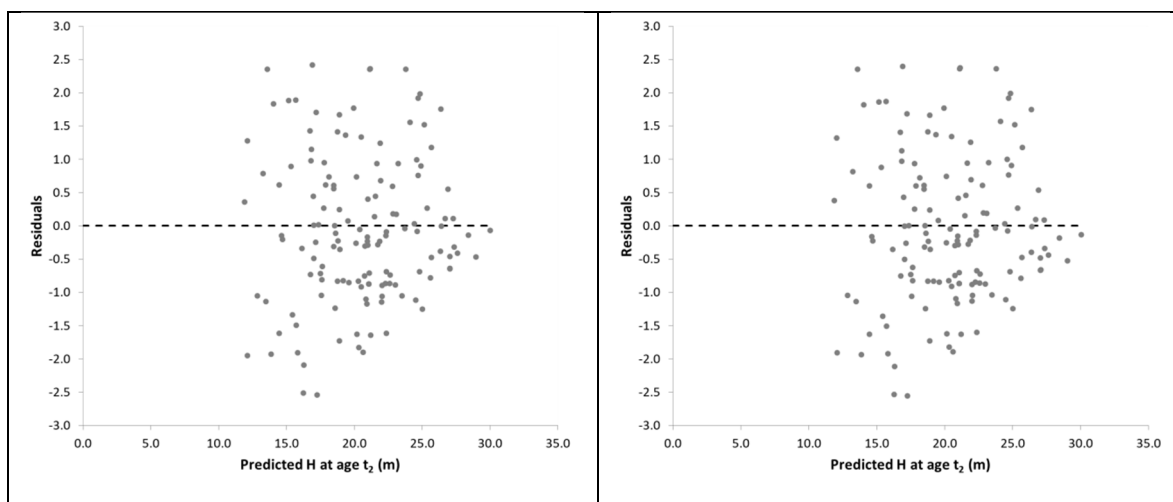
433

434 Table 3. Parameter estimates and goodness-of-fit statistics for the GADA equations fitted to  
 435 model dominant height growth using two different base-models: Von Bertalanffy (1957)  
 436 [E1] and Korf (1939) [E2].

Model	Parameter	Estimate	Approx. SE	RMSE(m)	MEF
E1	$b_1$	0.01829	0.0059		
	$b_2$	2.7972	1.0629	1.2366	0.9128
	$X$	3.6997	0.2847		
E2	$b_1$	33.1988	4.2729		
	$b_2$	0.4995	0.0612	1.2367	0.9128
	$X$	4.3487	1.3047		

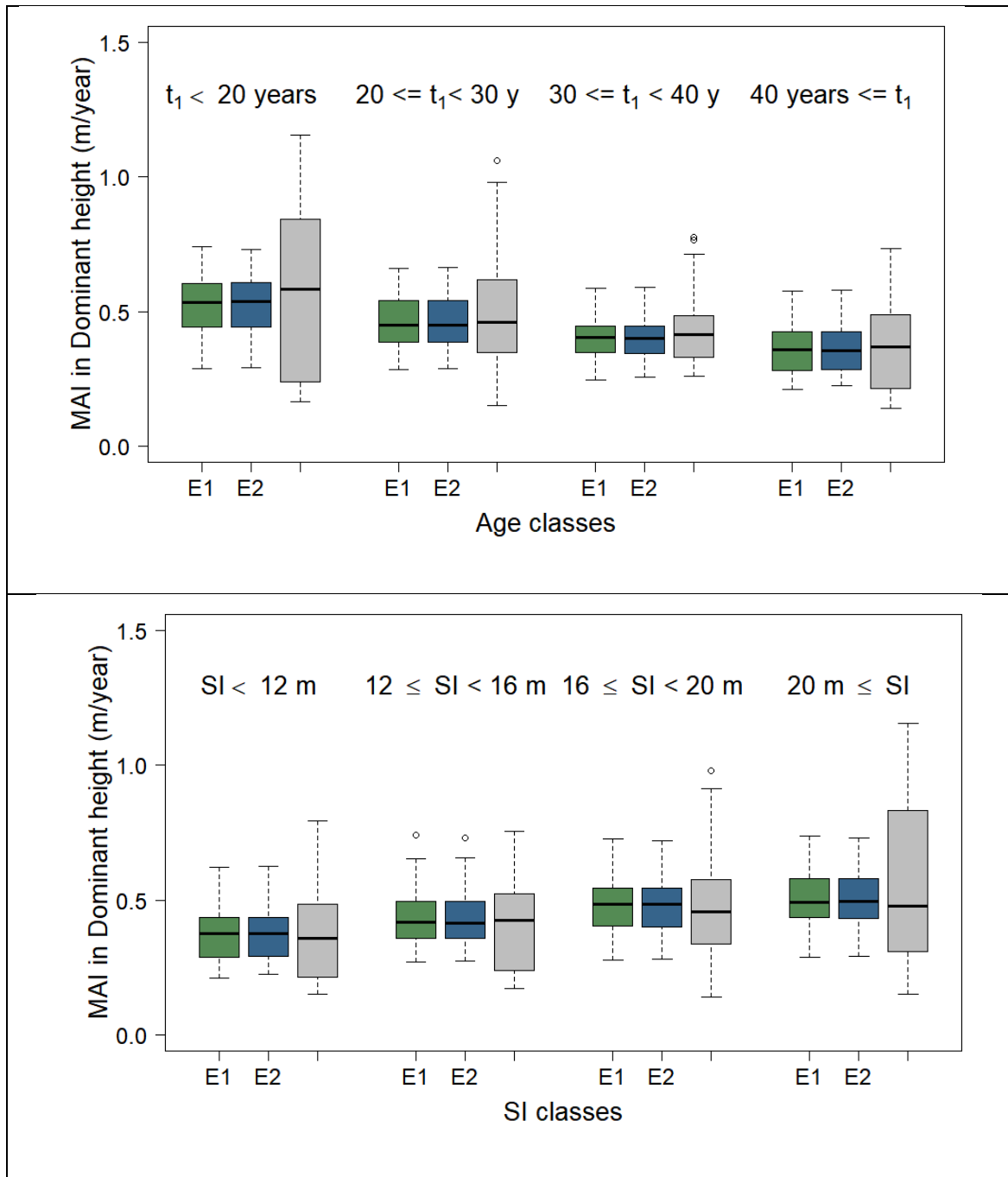
437 where,  $b_1$ ,  $b_2$  = global parameters,  $X$  = site-specific parameter

438



439 Figure 3. Plot of residuals versus predicted dominant height at age  $t_2$  from model E1 (left)  
 440 and E2 (right)

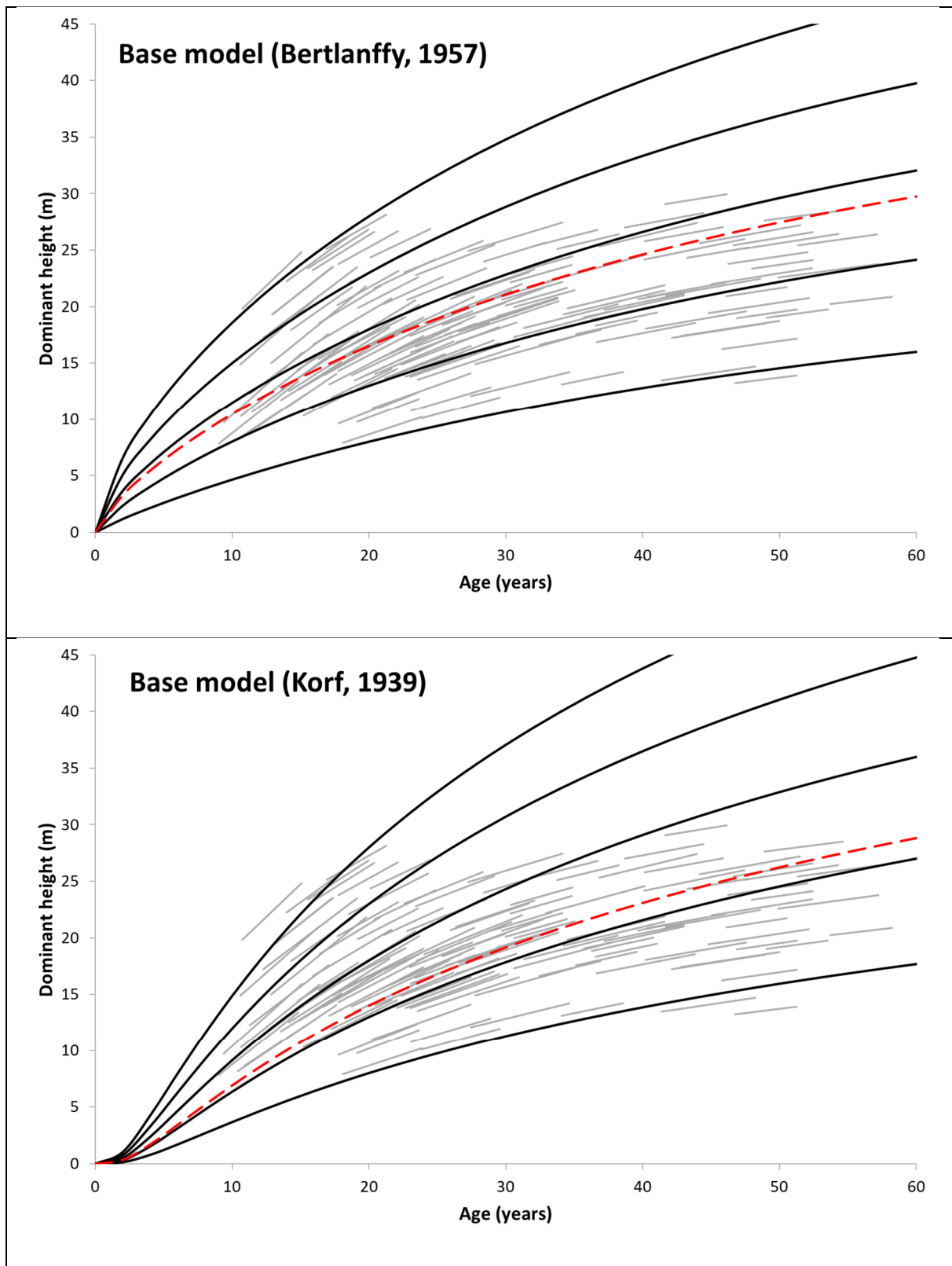
441



442 Figure 4. Boxplots of mean annual increments (MAI) in dominant height obtained with  
 443 each model ([E1] and [E2]) and field-observed values (grey colour) of the sample plots for  
 444 different age (upper) and site index classes (lower).

445

446 As shown in Figure 4, there are no perceptible differences between the results obtained  
 447 with both GADA equations analysed, so finally, to select the best model, a set of curves  
 448 corresponding to *SI*s of 8, 13, 18, 23 and 28 m at a reference age of 20 years was obtained  
 449 for each GADA equation and overlaid on the trajectories of the *H* of each sample plot for a  
 450 graphical analysis (Figure 5).



451 Figure 5. Height development curves (black lines) for site indices of 8, 13, 18, 23 and 28 m  
 452 at a reference age of 20 years, derived from equation [E1] (upper) and equation [E2]  
 453 (lower), overlaid on the trajectories of dominant height (grey lines) estimated for each  
 454 sample plot from the 99th height percentile from ALS data in 2009-2011 and 2015-2017.  
 455

456 Graphical inspection indicated that model [E1] slightly underestimate the growth rate  
 457 for young stands ranging from 10 to 20 years old, whereas model [E2] derived from the  
 458 base model proposed by Korf (1939), overestimated height growth for the best site  
 459 qualities, and the curves generated from this model seemed to increase more quickly than  
 460 the trajectories of sample plots. Overall, although both models have a very similar  
 461 behaviour, based on the superimposition of the site quality curves on the observed trends of  
 462 the plots, the age-independent equation [E1], derived from the Bertalanffy's (1957) base  
 463 model, was proposed for modelling dominant height growth for *P. pinaster* stands in  
 464 Galicia:

$$H_2 = e^X \left( 1 - \exp \left( -0.01829 \left[ \Delta t - \log \left( 1 - \left( \frac{H_1}{e^X} \right)^{X/2.7972} \right) / 0.01829 \right] \right) \right)^{2.7972/X} \quad [8]$$

465 where  $H_1$  and  $H_2$  are the dominant heights estimated using data from two ALS flight  
 466 surveys carried out for a time interval  $\Delta t$ , and  $X$  is the site-specific (local) parameter to be  
 467 estimated.

468 Once parameter  $X$  was obtained for a sample plot, the specific curve for that sample  
 469 plot can be obtained using the original base model equation with the global parameters  
 470 (equation [9]). The  $SI$  of the sample plot will be the dominant height estimated using the  
 471 specific equation for a reference age of 20 years (equation [10]):

$$H = \exp(X) \left( 1 - \exp(-0.01829t) \right)^{2.7972/X} \quad [9]$$

$$SI = \exp(X) \left( 1 - \exp(-0.01829 \cdot 20) \right)^{2.7972/X} \quad [10]$$

472 where  $H$  is the predicted dominant height (m) at age  $t$  (years),  $SI$  is the site index (m) and  $X$   
 473 is the site-specific (local) parameter.

474 The values of  $X$  parameter can be obtained from equation [8] using two dominant  
 475 heights ( $H_1$  and  $H_2$ ) estimated using data from two ALS flight surveys carried out for a time  
 476 interval  $\Delta t$ :

$$H_2 - e^X \left( 1 - \exp \left( -0.01829 \left[ \Delta t - \log \left( 1 - \left( \frac{H_1}{e^X} \right)^{X/2.7972} \right) / 0.01829 \right] \right) \right)^{2.7972/X} = 0$$

478 As discussed previously, it is necessary to restrict the estimates of parameter  $X$   
 479 obtained from the above equation to avoid unusual values. Solving equation [10] for  
 480 parameter  $X$  using the range of  $SI$  observed for the species in the study area (6.7 and 29.1  
 481 m; Álvarez-González et al., 2004) yielded estimates of 3.0037 and 4.1652. Therefore, a

482 minimum value of 3 and a maximum value of 4.1 were assumed for the site-specific  
 483 parameter.

484

485 *Site-specific parameter model*

486 The value of the site-specific parameter  $X$  required for practical use of equations 9 and  
 487 10 has some restrictions as it affects both the value of the asymptote and the growth pattern  
 488 of the curves generated with the GADA model. Taking into account a minimum value of 3  
 489 assumed for this parameter, the following model formulation was selected to guarantee  
 490 fulfilment of this limitation:

$$X = 3 * (1 + \exp(\alpha_0 + \alpha_i Y_i)) \quad [11]$$

491 where  $Y_i$  is a set of environmental variables and  $\alpha_0, \alpha_i$  are parameters to be estimated.

492 Equation [11] was log-linearized ( $\log(X/3 - 1) = \alpha_0 + \alpha_i Y_i$ ), and the MARS model  
 493 was fitted to the values of the new dependent variable.

494 The best results in terms of the generalized cross-validation (GCV) criterion was  
 495 obtained with a total of 14 basis functions (intercept, 5 first degree basis functions and 8  
 496 second degree basis functions) including 13 environmental variables. The expression for  
 497 the basis functions is shown in Table 4.

498

499 Table 4. Parameter estimates and basis function of the MARS model fitted to estimate the  
 500 site-specific parameter  $X$  from environmental variables.

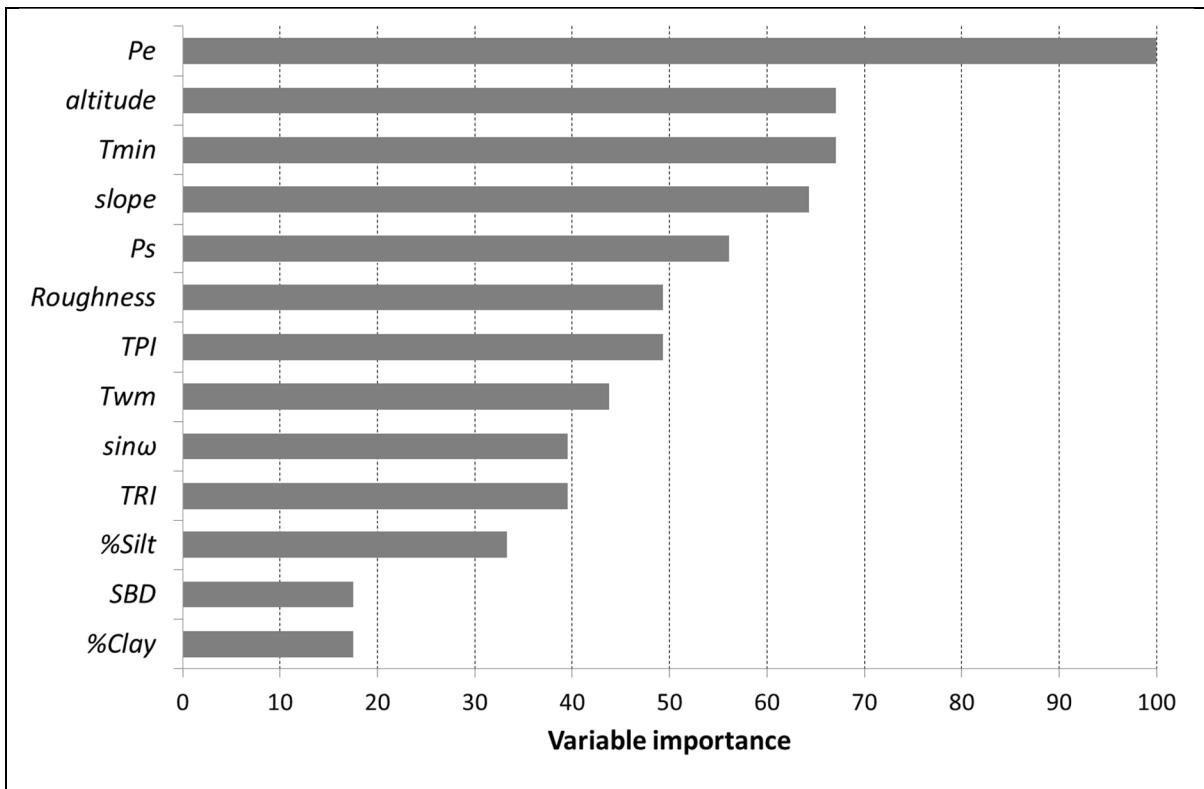
Parameter	Estimate	Basis function
$\alpha_0$	-1.1774	
$\alpha_1$	1.7070	(6.0695- $T_{min}$ )
$\alpha_2$	-0.0675	(%Clay-13.8485)
$\alpha_3$	-5.9814	(SBD-1.0824)
$\alpha_4$	-0.0155	(altitude-533.5570)
$\alpha_5$	-0.0161	(869.2190- $Pe$ )
$\alpha_6$	0.0001	(altitude-533.557)· $Ps$
$\alpha_7$	0.0112	( $T_{min}$ -6.0695)·(15.9403- $slope$ )
$\alpha_8$	0.0171	( $T_{min}$ -6.0695)·(884.5470- $Pe$ )
$\alpha_9$	-0.0001	(30.2877-%Silt)·(533.557-altitude)
$\alpha_{10}$	0.0437	(SBD-1.1251)·(533.557-altitude)
$\alpha_{11}$	-0.1129	(2.2205- $TRI$ )·(0.6824- $\sin\omega$ )

$\alpha_{12}$	-0.0992	$(0.1424-TPI) \cdot (Roughness-6.7759)$
$\alpha_{13}$	-0.0019	$(19.8899-Twm) \cdot (Pe-869.2190)$

501

502 The parameters and the basis functions of the MARS fitted model were substituted in  
 503 equation [11], and the  $X$  estimates obtained with this transformed model explained 57.69%  
 504 of the observed variability in the site-specific parameter, with a GCV value of 0.0229.

505 The importance of each environmental variable included in the MARS model was  
 506 evaluated in terms of the decrease in the residual sum of squares of the model due to its  
 507 inclusion, and the results were scaled so that the variable yielding the largest summed  
 508 decrease is assigned a value of 100. Figure 6 shows the scaled importance of the  
 509 environmental variables included in the MARS model.

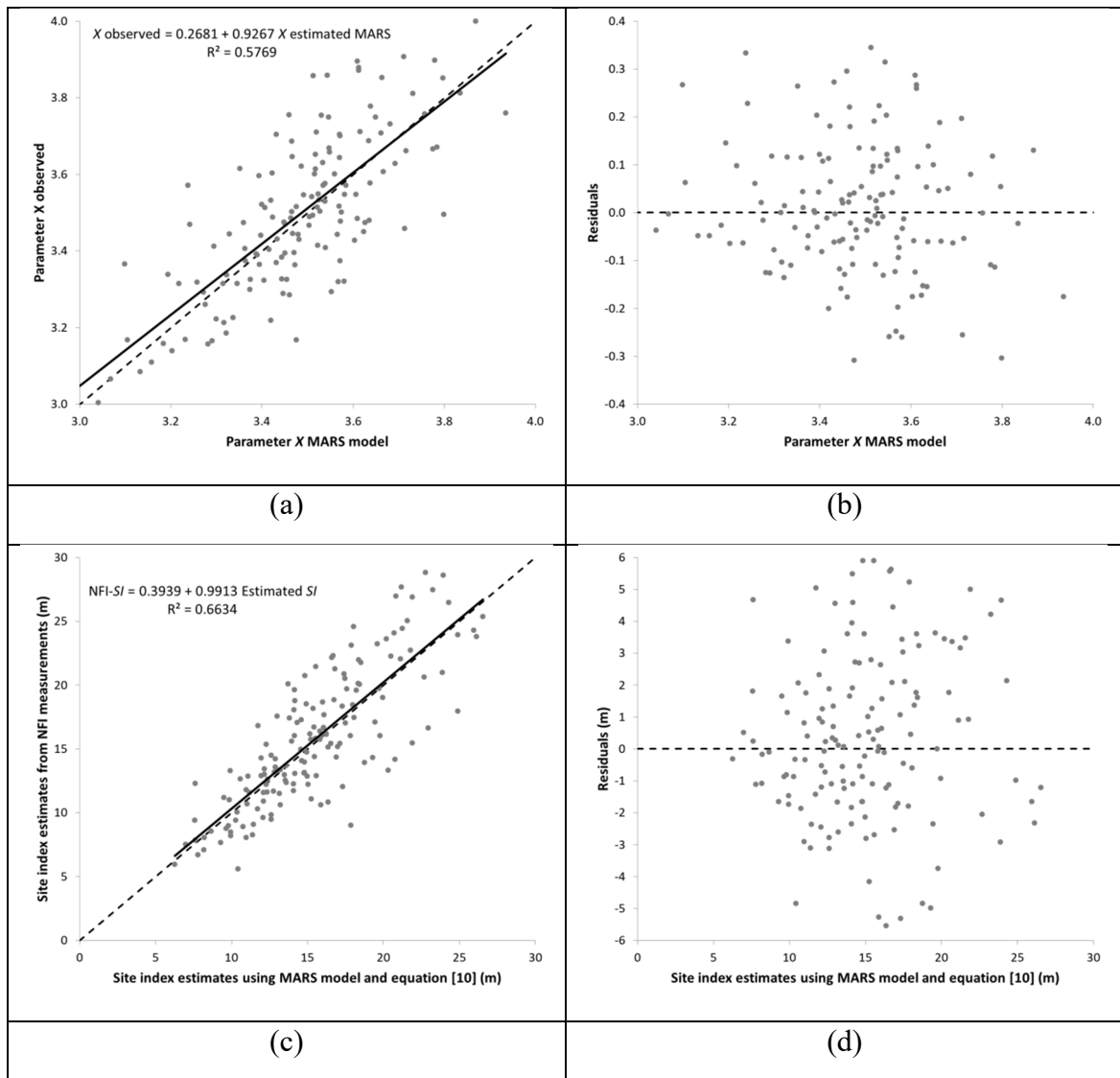


510 Figure 6. Scaled importance of the environmental variables included in the MARS model in  
 511 terms of residual sum of square reduction (the variables are described in Table 2)

512

513 To assess the performance of the MARS model fitted, four different plots were  
 514 analysed. In a first step, the accuracy of estimates of the MARS model was evaluated by  
 515 analysing the plots of site-specific parameter observed versus site-specific parameter  
 516 estimated with the model (Figure 7a) and residuals versus predicted values of the site-  
 517 specific parameter obtained with the model (Figure 7b). Finally, the  $X$  estimates were used

518 to obtain the  $SI$  of each sample plot using equation [10] and plots of observed versus  
 519 predicted  $SI$  (Figure 7c) and residuals versus predicted  $SI$  (Figure 7d) were analysed.  
 520 .



521  
 522 Figure 7. Scatter plot of: a) observed site-specific  $X$  parameter versus site-specific  $X$   
 523 parameter estimated using the MARS model; b) residuals versus predicted values of local  
 524 parameter ( $X$ ) estimated with the MARS model; c) observed  $SI$  values versus predicted  $SI$   
 525 values using the MARS model to estimate the  $X$  parameter and then equation [10]; d)  
 526 residuals versus predicted values of  $SI$  estimated using the MARS model and then equation  
 527 [10] (d). The solid line represents the fitted linear model.  
 528

529 **Discussion**

530 Galicia is one of the most important regions in Spain in terms of forestry production  
531 (69% of the land is covered by forest), and pure *P. pinaster* stands, the most extensive  
532 forest type in Galicia until the end of the 20th century, still represent more than 15% of the  
533 Galician forests, i.e. they are the second most abundant productive forests in Galicia  
534 (MAPA, 2018). Obtaining further knowledge about forest ecosystems could improve forest  
535 management and decision-making for improved productivity and competitiveness of the  
536 forestry sector in Galicia.

537 The present study demonstrates the potential of bi-temporal ALS data, which when  
538 combined with other sources of freely available environmental data enabled us to develop  
539 regional *SI* models for *P. pinaster* stands in Galicia. The multitemporal low-density ALS  
540 data from the PNOA (with a planned temporal resolution of 6 years) is promising auxiliary  
541 data for improving traditional *SI* modelling, while spatially capturing changes across forest  
542 landscapes. The method developed, which combines two ALS surveys and the SNFI  
543 ground data sets, enables shifts from age-dependent to age-independent *SI* models, while  
544 respecting the underlying principles of *SI* modelling (see Introduction). Thus, the age-  
545 independent models presented boost the capacity of ALS to account for changes over time  
546 without any constraints to age measurement. Noordermeer et al. (2020) and Socha et al.  
547 (2019, 2020) recently highlighted the potential of multitemporal ALS data observations for  
548 modelling *SI*. The proposed methods could be applied to mapping local site-specific *SI*  
549 models, to address the problem of uncertainty in forest growth estimation due to local  
550 variation and also to improve large-scale forest productivity estimation in regions such as  
551 Galicia.

552 The base models used to develop the GADA equations [E1] and [E2] have been widely  
553 used to develop age-dependent site index curves. For example, Betarlanffy's base model  
554 was selected in the studies by Scolforo et al. (2016), for eucalypt plantations in Brazil, and  
555 by Seki and Sakici (2017), for Crimean pine in Turkey, whereas Korf's base model proved  
556 more accurate for even-aged natural longleaf pine in US (Lauer and Kush, 2010) and for  
557 Calabrian pine in Turkey (Kahriman et al., 2018) than Bertalanffy's base model. Although  
558 both GADA equations [E1] and [E2] are quite flexible and proved equally accurate for  
559 estimating dominant height, an ideal solution was not achieved since [E1] model tended to  
560 underestimate the height growth for young stands and [E2] model tended to overestimate  
561 for high qualities. Nonetheless, it is usually recommended to apply with caution the site

562 quality curves to young stands due to possible errors in dominant height growth estimates  
563 and site indices (e.g. Weiskittel et al., 2011), so we considered that the behaviour of the  
564 [E1] model, derived from the Bertalanffy's (1957) base model, was the most appropriate to  
565 model the growth in dominant height.

566 The selected GADA equation to estimate the height growth curves using multitemporal  
567 ALS data explained more than 91% of dominant height growth variability. This value is  
568 lower than that reported by Socha et al. (2020), which explained the 99% of the observed  
569 variability using a 99th percentile, 30-m resolution layer for Scots pine in Poland. However,  
570 the model developed in the present study is age-independent, while that developed by  
571 Socha et al. (2020) is age-dependent, so that the comparison is not conclusive.

572 A key factor in the development and practical use of the fitted GADA model is an  
573 accurate estimation of the dominant height from LiDAR metrics. The adjusted linear  
574 relationship in this study explains just over 89% of the observed variability using the 99th  
575 percentile as the only independent variable (Figure 2). This result contrasts with that  
576 obtained by Socha et al. (2020) in a study on the effect of different LiDAR metrics on site  
577 index model estimates. These authors observed that the use of the 99th percentile tended to  
578 underestimate the growth in dominant height, especially in older stands with a reduced  
579 number of trees and broken crown closures, and recommended the use of the 100th  
580 percentile or the estimation of the dominant height by using individual tree detection.  
581 Although the model obtained in this study based on the 99th percentile has a slight bias (-  
582 0.23 m), the use of the 100th percentile resulted in more biased (-1.33 m) and less accurate  
583 estimates (RMSE increased from 1.74 to 2.50 m). One of the possible reasons for this  
584 difference could be the definition of dominant height used in each case, the mean height of  
585 the 100 thickest trees per hectare in our case and that proposed by Rennolls (1978) in Socha  
586 et al. (2020). Another factor that could cause these contradictory results is the high density  
587 of *Pinus pinaster* stands in the study area with a mean value of 669 trees ha<sup>-1</sup>, so the  
588 situations where Socha et al. (2020) found higher underestimates (low-density stands with  
589 open coverages) are not so frequent. In addition, another factor possibly influencing the  
590 results is the low density of the ALS data in our study (0.5-1 pulses m<sup>-2</sup>) compared with that  
591 used by Socha et al. (2020) (7-10 pulses m<sup>-2</sup>).

592 Pulse density is widely recognized to be one of the most important factors influencing  
593 the accurate estimation of tree height from ALS (Hyypä et al., 2008; Roussel et al., 2017).  
594 It has been established that LiDAR systems tend to underestimate tree height (Hyypä et al.  
595 2008), which is consequence of the operation of the scanner and the inadequate

596 representation of canopy apices due to low point density (e.g. Gonçalves-Seco et al., 2011;  
597 Roussel et al., 2017). Laser pulse density and vegetation structure are also important issues  
598 in relation to the height accuracy of the laser-derived Digital Elevation Model (DEM) to  
599 normalize the height (Clark et al., 2004; Raber et al., 2002; Valbuena et al., 2011),  
600 especially on steep terrain (Estornell et al., 2011). Therefore, the determination of the  
601 height increment working with a low-density point cloud, especially in the short term  
602 interval, could be subject to a large error (Hopkinson et al., 2008). According to Hopkinson  
603 et al. (2008), the ALS observation interval in our study was considered enough to capture  
604 height growth trend without large errors. These authors, in a study of height growth in red  
605 pine conifer plantations using low-density ALS data, reported that both the absolute and  
606 relative magnitudes of plot-level uncertainty decrease in a power–function relationship with  
607 the annual time increment, such that after three years the level of uncertainty in the LiDAR  
608 growth estimate rapidly drops to an operationally acceptable level of approximately 10%;  
609 with additional time providing minimal increases in accuracy (Hopkinson et al., 2008). This  
610 is shorter than a typical five-year time interval of field-based repeat height assessment for  
611 similar conifer plantations.

612 Another limitation of the proposed approach is that two ALS flights are needed to  
613 produce the  $\Delta H$  curve of a new sample plot or stand to estimate the site-specific parameter  
614 ( $X$ ). Thus, a compatible model for predicting and projecting height growth and including a  
615 MARS equation to estimate the site-specific local parameter  $X$  from environmental  
616 variables was developed for the species in the study region. The fitted MARS equation  
617 explained more than 57% of the total variation in the observed site-specific parameter based  
618 on physiographical, edaphic and climate variables (Figure 7a). Moreover, the site indices  
619 obtained from estimates of parameter  $X$  using equation [10] explained more than 66% of  
620 the observed variability in site indices derived from field measurements (Figure 7c).

621 According to the results of the variable importance analysis (Figure 6), the  
622 geomorphometric variable (*elevation*) and potential evapotranspiration ( $Pe$ ) were the most  
623 important predictors for describing the relationship between site-specific parameters and  
624 environmental variables. These results also suggest that the physiographical variables  
625 extracted from DEMs may be just as important or more important than edaphic and climatic  
626 variables for characterizing the potential production of *P. pinaster* in the study area, as in  
627 addition to *elevation*, the variables *slope*, *TPI* and *Roughness* were among the subsequently  
628 most important variables for site-specific parameter modelling. This is consistent with the  
629 findings of Rodríguez-Soalleiro (1995) and Álvarez-Álvarez et al. (2011), who used

630 elevation as a significant variable for *P. pinaster* *SI* modelling in Galicia and Asturias,  
631 respectively. *TRI* and *sin $\omega$*  were also included among the 13 significant environmental  
632 variables in the MARS model, highlighting the importance of small-scale orographic  
633 effects in Galicia. It is important to note that *slope*, *TRI*, *sin $\omega$*  were also included as  
634 physiographic predictors in the MARS model developed by González-Rodríguez and  
635 Diéguez-Aranda (2020) to estimate the *SI* of *P. radiata* plantations, which used 13  
636 environmental variables and a 200-m resolution DEM for a regional study in Galicia.

637 Four climatic variables were also included in the fitted MARS equation: more  
638 specifically, *Pe* (Potential evapotranspiration), *Tmin* (mean annual minimum temperature)  
639 and *Ps* (Mean precipitation in summer) were among the 5 most important variables  
640 explaining the variation in the site-specific parameter *X*. Similar results have been obtained  
641 for the same species in the neighbouring region of Asturias (NW Spain) (Álvarez-Álvarez  
642 et al. (2011) and Arias-Rodil et al. (2015)). The first researchers evaluated the effects of  
643 foliar nutrients and environmental factors on site quality of *P. pinaster* stands in Asturias,  
644 observing that *SI* was higher in stands with a higher mean summer temperature (*Ts* in °C)  
645 and lower winter precipitation (*Pw*, in °C). By contrast, Arias-Rodil et al., (2015) fitted  
646 different linear relationships to relate a site-specific parameter of a dominant height growth  
647 model and environmental variables, obtaining the best results with minimum mean  
648 temperature of the coldest month (*Tcm*) and mean annual precipitation (*P*) as independent  
649 variables. The models developed for Asturias (Álvarez-Álvarez et al., 2011; Arias-Rodil et  
650 al., 2015) were based on sample plot grids established in a pure Atlantic climate in the area  
651 of distribution of *P. pinaster*, while the sample plots used in this study were spread along  
652 areas characterized by Atlantic, continental and even Mediterranean climate conditions,  
653 which may explain the relative importance of *Ps* and *Tmin*.

654 In the present study, *Pe* was the most important variable in the MARS model in  
655 terms of reduction of the residual sum of squares. Numerous previous studies have reported  
656 the influence of *Pe* on forest growth (e.g. Kimmins et al., 1990; Wickramasinghe, 1988;  
657 Martínez-Vilalta et al., 2008; Liu and El-Kassaby, 2018). This variable is closely related to  
658 water deficit (Gracia et al., 1999) and worsening of drought conditions (Klos et al., 2009;  
659 Scolforo et al., 2013; Vicente-Serrano et al., 2014; Gazol et al., 2017; Gleason et al., 2017),  
660 both of which can constrain forest growth. In agreement with other studies in eucalyptus  
661 and pine stands in Brazil (Scolforo et al., 2013; Campoe et al., 2016; Scolforo et al., 2020),  
662 our findings suggest that the inclusion of *Pe* as predictor could integrate the effect of  
663 physiological, soil and climate variables in dominant height growth modelling, thereby

664 increasing the ability of the model to estimate site quality. The *SI* model proposed by  
665 Scolforo et al. (2020) included mean annual soil water deficit ( $\text{mm year}^{-1}$ ), which used the  
666 monthly differences between potential and actual evapotranspiration, calculated by the  
667 water balance model of Thornthwaite-Mather. Campoe et al. (2016) found that *Pe*, along  
668 with other variables such as maximum temperature and vapour deficit, were related to the  
669 annual increase in tree basal area in pine and eucalyptus plantations in Brazil. Bravo-  
670 Oviedo et al. (2008) indicated that drought length, mean annual temperature, and  
671 precipitation regimes prior to the growth season (total autumn and winter precipitation)  
672 were considered regional differentiation factors to explain differences in the dominant  
673 growth pattern of *P. pinaster* stands for four different regions in Spain. Finally, Condés and  
674 García-Robredo (2012) reported the influence of precipitation, temperature, Emberger  
675 index and free climatic intensity (the latter dependent on *Pe*) on the stand volume growth in  
676 *P. halepensis* stands in SE Spain.

677 The edaphic variables silt (*%Silt*), clay (*%Clay*) and soil bulk density (*SBD*,  
678  $\text{Mg m}^{-3}$ ), which are related to soil water capacity, were the most important edaphic drivers  
679 explaining *P. pinaster* height growth. However, the strictly edaphic variables were the least  
680 important from the point of view of reducing the residual sum of squares in the MARS  
681 model. This is also consistent with previous findings of Arias-Rodil et al. (2015), who  
682 reported that *SI* models including a third variable related to soil characteristics explained  
683 only 1–2 % more variability than models including only climate variables.

684 We developed a two-equation system to predict  $\Delta H$  and *SI*, which represents an  
685 advancement towards improving forest management decisions for *P. pinaster* stands in  
686 Galicia and identifying remotely sensed features that enable reliable estimation of  $\Delta H$  and  
687 *SI*. The system is flexible, and it supports forest management planning, maximizing the use  
688 of ALS surveys and ground data from SNFIs. The value added by this research is related to  
689 the opportunity to generate *SI* models for application at several spatial scales, for a wide  
690 geographic distribution of the forest species at regional level and supporting the change in  
691 paradigm in the field of forest growth and site quality (age independency).

692 With the rapid development of forestry-oriented enterprise consultancies and the  
693 solid progress of the PNOA-LiDAR project, ALS surveying is becoming a multi-purpose  
694 tool in Spain. In addition, new accurate positioning of SNFI plots (GNSS-based) will  
695 contribute greatly to improving large-scale estimates of biophysical forest attributes  
696 (Guerra-Hernández et al., 2016, 2019). For instance, the models presented here can be used  
697 for automatic mapping of *SI* over large areas of *P. pinaster* stands, representing forest

698 productivity, which may be of interest for inclusion in the Spanish Forest Map (SFM). The  
699 first countrywide ALS coverage is already available for the whole Spanish territory, and the  
700 second countrywide ALS coverage will soon be available for all Spanish regions, with the  
701 theoretical average point density data ranging from 0.5 to 14 points m<sup>-2</sup>. Therefore, in the  
702 near future, the amount of data available from ALS surveys will increase the importance of  
703 change detection by ALS in forest growth modelling (Yu et al., 2006; Hopkinson et al.,  
704 2008; Socha et al., 2017). Finally, Ice-Sat2 and GEDI NASA's missions could also  
705 improve forest productivity modelling by increasing the availability of data in real time and  
706 decreasing the temporal resolution. For future improvement of cartography, better  
707 resolution of edaphic and climate rasters should also be considered for assessing *SI* and  
708 forest growth by using environmental variables.

709

## 710 **Conclusions**

711 SNFI data and multitemporal low density ALS data from the PNOA were used to  
712 develop a regional site index model for *P. pinaster* stands in Galicia. A two-equations  
713 system comprising one equation from two ALS flights and a complementary equation  
714 including environmental variables as predictors (for when only one ALS flight is available)  
715 is proposed. Consequently, site quality maps for the whole region could easily be  
716 developed by combining the age-independent GADA model proposed, information from  
717 any future ALS flight surveys and cartography of the distribution of the species in Galicia.  
718 The MARS model developed in order to estimate the site-specific parameter of the GADA  
719 model from environmental variables performed well, with a relatively robust and  
720 parsimonious form (13 predictors). The models presented here can be used to map *SI* over  
721 large areas of forest automatically, determining forest productivity at a much finer spatial  
722 resolution than usually possible operational inventories. Future challenges are related to  
723 improving the resolution of the raster models, which are currently limited by the spatial  
724 sampling of climate and edaphic variables.

725

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738

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740 Conceptualization, JG, SA and JGA; Methodology, JGA; Data preparation, JG, SA, EGF,  
741 ADR and JGA; Data Analysis, JGA; Research, JG, EGF, SA, AP and JGA; Resources, VS  
742 and JGA; Writing—Original Draft Preparation, JG, SA, EGF, AP, ADR and JGA;  
743 Writing—Review & Editing; JG, SA, EGF, AP, ADR and JGA and; Project  
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752

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